

## Individual chlorinated biphenyls and pesticides in tissues of some cetacean species from the North Sea and the Atlantic Ocean; tissue distribution and biotransformation

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### Summary

Individual chlorinated biphenyls (CBs) and organochlorine pesticides ( $\alpha$ - and  $\gamma$ -HCH, dieldrin, HCB, p,p'-DDT, p,p'-DDD, and p,p'-DDE) were investigated in blubber, brain, heart, kidney, liver and muscle of 19 cetaceans from the Dutch coastal area, the open North Sea and the western Atlantic. Multidimensional gas chromatography (MDGC) with electron capture detection (ECD) was used to check the results of single capillary column GC-ECD data.

Absolute concentrations and concentration patterns of individual compounds were investigated to describe their distribution between tissues and to interpret the findings in terms of lipophilicity and persistence to biotransformation of individual compounds. A clear relation could be established between the chemical structure of each CB and its biomagnification in marine mammal tissues. Persistent congeners are biomagnified but metabolizable congeners appeared at much lower extractable lipid based concentrations than in the food sources. The CBs with the highest concentrations in the mammal tissues were hexa- and heptachloro-biphenyls.

The results are compared with published data on individual CBs in cetaceans and in blood of seals from the coastal area. CB patterns in different tissues of any organism were almost identical. A large similarity in the patterns of persistent congeners was also found between tissues and organs of mammals from widely separated oceanic areas, although absolute concentrations differed sometimes by more than an order of magnitude. Differences were found in the relative contributions of less persistent congeners. The apparent rates of biotransformation were higher in seals than in cetaceans, for which the lowest rates were generally observed in the animals from the open sea and in particular in *Physeter macrocephalus* and also *Sotalia fluviatilis*.

Especially CB congeners with vicinal H-atoms only in the m,p positions were more rapidly metabolized in seals than in cetaceans. This can be due to the lower activity of the MFO enzyme system at lower CB levels or to physiological differences between species. CB congeners with vicinal H-atoms in the o,m positions with not more than one ortho-chlorine were metabolized in each case.

Of the most toxic CB congeners, i.e., those showing a 3-MC or mixed 3-MC and PB- type MFO induction in rats, CBs -77, -105, -118, -128, and -138 were present (although at low concentration levels), but CBs -81, -114, -123, -126, -167, -157, -169, and -189 were below detection levels.

The concentrations of individual CBs were of similar magnitude as those of some pesticides. In Wadden Sea fish and in porpoises from the coastal North Sea, p,p'-DDD was the dominant contributor to  $\Sigma$ -DDT but in the open North Sea and the western Atlantic, p,p'-DDE was by far the largest contributor (80-90%).

### Introduction

Several organochlorines such as Polychlorinated biphenyls (PCBs) and pesticides, e.g., members of the DDT family, dieldrin, etc., are strongly lipophilic compounds, which concentrate in lipid tissues of marine organisms. There is ample evidence that 'PCBs' and 'DDTs' are at least partly responsible for reproductive and immunological abnormalities (Hook & Johnels, 1972; Fuller & Hobson, 1986; Helle, Olsson & Jensen, 1976a,b; Olsson, Johnels & Vaz, 1975; Anon., 1987) observed in some marine mammal populations. This has been reported for sea lions in Californian coastal waters (De Long *et al.*, 1973; Gilmartin *et al.*, 1976), for common seals and harbour porpoises of the Dutch Wadden Sea (Reijnders, 1978, 1979, 1986), for grey seals, ringed seals and harbour seals in the Baltic (Helle, 1980, 1981, 1986; Helle, Olsson & Jensen, 1976a,b; Almkvist, 1982) and for belugas in the St. Lawrence

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river estuary (Béland, Michaud & Martineau, 1987; Martineau *et al.*, 1987).

In fish and invertebrates, equilibrium partitioning between body lipids, blood and ambient water is the dominant mechanism for elimination of apolar compounds (Hamelink, Waybrant & Ball, 1971). This mechanism, involving exchange between blood and water via the gills is not available for marine mammals. This accounts at least partially for the considerably higher concentration levels of some apolar compounds in marine mammals than in lower organisms.

The relative contributions of individual chlorinated biphenyls (CBs) to their mixtures (i.e., the CB patterns) in marine mammal tissues differ from those in lower organisms. In particular, congeners with vicinal H-atoms in the m,p positions or with vicinal H-atoms in the o,m positions in combination with not more than one ortho-chlorine, are missing in mammals or present at only minor concentration levels. This is an indication for the significant role of enzyme-mediated metabolism in the elimination of these congeners that are transformed into more polar compounds with increased water solubility. Also reactivity may be increased, (Boon *et al.*, 1987). For example, lower plasma levels of vitamin A and thyroxine in rats and mice appeared to be caused by a metabolite of CB-77, but not by the parent compound (Brouwer *et al.*, 1986). Metabolism is of vital importance for the elimination of lipophilic substances in marine mammals (Abdel-Hamid, Moore & Matthews, 1981; Gage & Holm, 1979; Sipes *et al.*, 1980).

In a recent literature review (Reuthergaardh & Knap, 1987) it was concluded that (i) information on organochlorine levels in other tissues than blubber is scarce, (ii) much less information is available for cetaceans than for seals and (iii) practically all data have been obtained with the use of packed column gaschromatographic separation techniques with technical PCB mixtures (e.g., Aroclors) as reference materials. The resulting quantitative data on 'PCBs' contain therefore very limited useful information on individual chlorinated biphenyls. As the latter have different physical chemical properties, a more accurate insight into the mechanisms of uptake, distribution between tissues and elimination can be gained from studies of individual, well defined congeners in various tissues, including the major depot tissue (blubber) as well as possible target organs.

We report here on the distribution of chlorinated biphenyl congeners and organochlorine pesticides in blubber, brain, heart, kidney, liver and muscle of 19 Cetaceans originating from the Dutch coast, the open North Sea and the open and western Atlantic. Most animals were stranded harbour porpoises (*Phocoena phocoena*), whose population in coastal

North Sea waters has declined very strongly in the last decades (Verwey, 1975). The other animals were stranded open ocean animals or animals captured along the Western Atlantic coast, transferred to the Dolfinarium at Harderwijk in the Netherlands and fed with mackerel from the open Atlantic until death (when their tissues were sampled for analysis). The data will be used for comparison of absolute and relative concentrations of individual compounds between tissues and species from different regions in relation to food source and lipid physiology. Also, the present data on cetaceans are compared in detail with published data of CB congeners in harbour seals from the Dutch coastal area (Boon *et al.*, 1987).

## Methods and materials

### *Collection of animals*

Most animals described here were found dead along the Dutch coast in 1977-1979. Collection and transport of the carcasses of the protected species were authorized and controlled by two designated institutions (Zoological Museum, Amsterdam and Museum for Natural History, Leiden). Samples were taken under appropriate non-contaminating conditions, deep frozen in precleaned glass containers and transported to the Netherlands Institute for Sea Research at Texel. Only when degradation had not proceeded too far, were animals selected for analysis. Animals which had been caught alive were kept for some time at the Dolfinarium in Harderwijk. The available data for all animals is given in Table 1. Animal 15, a male sperm whale, was much larger than all other animals investigated.

### *Analytical Methods*

After thawing of samples at room temperature, sub-samples were taken, weighed and extracted with n-hexane in a Soxhlet extractor for 10 hours. Lipid content was determined as the n-hexane soluble fraction. An aliquot of the extract, containing 40-50 mg lipid was used for the analysis of organochlorines. Cleanup was performed over alumina and organochlorines were separated into classes over a silica column. Details have been described before (Duinker & Hillebrand, 1983b).

The first silica fraction contained hexachlorobenzene (HCB), p,p'-DDE and all CBs quantitatively. The second fraction contained  $\alpha$ - and  $\gamma$ -hexachlorocyclohexane (HCHs), dieldrin, p,p'-DDT, p,p'-DDD and several other, yet unidentified compounds. Analysis was performed by temperature programmed capillary gas chromatography with electron capture detection (GC-ECD) on a SE-54 column.

Compounds were identified with a mixture of pure reference compounds on the basis of retention times.

**Table 1.** Identification and specification of all mammals discussed in the paper

Animal number and species	Sex	Length (cm)	Weight (kg)	Age (y)	Found		Status
					Location	Time	
1 <i>Phocoena phocoena</i>	f	135	29	7	53°22N, 5°15E Dutch coast	7.8.78	dead
2 <i>Phocoena phocoena</i>	f	128	27	2.5	52°42N, 4°10E Dutch coast	29.3.78	dead
3 <i>Phocoena phocoena</i>	f	151	49	—	53°15N, 4°55E Dutch coast	12.4.79	dead
4 <i>Phocoena phocoena</i>	f	93	12	—	52°8N, 4°15E North Sea	7.3.77	dead
5 <i>Phocoena phocoena</i>	f	—	58	—	52°50N, 4°38E Dutch coast	11.6.78	alive <sup>a</sup>
6 <i>Phocoena phocoena</i>	f	155	39	7.5	52°60N, 4°40E Dutch coast	14.12.78	dead
7 <i>Phocoena phocoena</i>	f	153	43	>11	53°24N, 6°25E Dutch coast	8.3.78	dead
8 <i>Phocoena phocoena</i>	f	160	—	—	51°51N, 3°58E Dutch coast	17.10.78	dead
9 <i>Phocoena phocoena</i>	f	150	—	—	53°10N, 4°50E Dutch coast	5.4.77	dead
10 <i>Phocoena phocoena</i>	m	105	—	<1	52°7N, 4°15E Dutch coast	20. 3.77	dead
11 <i>Phocoena phocoena</i>	m	105	18	—	52°22N, 4°33E Dutch coast	14. 4.77	dead
12 <i>Lagenorhynchus albirostris</i>	f	242	100	—	57°N, 1°W North Sea	4.7.77	dead
13 <i>Lagenorhynchus albirostris</i>	m	239	134	—	57°N, 1°W North Sea	4.7.77	dead
14 <i>Lagenorhynchus albirostris</i>	f	235	95	—	53°7N, 4°45E North Sea	30.9.77	dead
15 <i>Physeter macrocephalus</i>	m	1522	30 000	11	52°34N, 4°35E Dutch coast	15.12.79	dead
16 <i>Delphinus delphis</i>	f	170	—	—	53°10N, 4°50E Dutch coast	6.3.79	dead
17 <i>Sotalia fluviatilis</i>	f	170	50	8	9°25N, 75°40W San Antero, Cordoba	10.7.77	alive <sup>b</sup>
18 <i>Sotalia fluviatilis</i>	m	170	55	9	9°25N, 75°40W San Antero, Cordoba	10.7.77	alive <sup>c</sup>
19 <i>Tursiops truncatus</i>	f	237	—	32	26°6N, 79°55W Off Fort Lauderdale	3.7.65	alive <sup>d</sup>

a–d: animals were transferred to Dolfinarium, Harderwijk, the Netherlands, where they died on 7.8.78. (a), 14.1.78 (b), 19.12.77 (c, named Ramon) and 7.1.78 (d, named Mama). Registration numbers at the Zoologisch Museum, Amsterdam (ZMA) and Rijks Museum voor Natuurlijke Historie (RMNH, Leiden) are ZMA 20013 (1), ZMA 19799 (2), ZMA 19205 (3), ZMA 19203 (4), ZMA 20164 (5), ZMA 20336 (6), ZMA 19797 (7), RMNH 27281 (8), RMNH 25754 (10), ZMA 19206 (11), Museum den Haag (15), ZMA 19780 (17) and ZMA 19775 (18).

This did not resolve the problems associated with the possibility of the presence of co-eluting compounds. It was realized that this was the case for several peaks on the basis of peak shape and it was suspected to be the case for in principle each PCB peak (Duinker & Hillebrand, 1983a). This problem has been resolved since Mullin *et al.* (1984) have published the retention properties of all 209 congeners on a SE-54 column. In addition, multidimensional gas

chromatography with electron capture detection (MDGC-ECD) was introduced recently. The technique allows the unambiguous determination of individual CBs and other compounds in complex mixtures, e.g., technical PCB mixtures and seal blubber (Duinker, Schulz & Petrick, 1988a; Schulz, Petrick & Duinker, 1989). Thus, the presence of the following pesticides was recently confirmed for the present samples at the Institute for Marine Research

at Kiel: p,p'-DDT, p,p'-DDD, p,p'-DDE, HCB, dieldrin,  $\alpha$ -HCH, and  $\gamma$ -HCH. The technique showed that aldrin, endrin and o,p'-DDD were not present, although the retention times on SE-54 were consistent with their presence (Fig. 1). In the remainder of this paper, p,p'-DDD, p,p'-DDT and p,p'-DDE will be indicated by DDD, DDT and DDE.

The CBs often found to be present in relatively high concentrations in the mammal tissue extracts as well separated peaks (identification in terms of numbers) are: 52, 49, 44 (tetrachloro-CBs), 101, 118 (pentachloro-CBs), 149, 153, 138 (hexachloro-CBs), 187, 183, 177, 172, 180 (heptachloro-CBs), 201, 194 (octachloro-CBs), 206 (nonachloro-CB), and 209 (decachloro-CB). The molecular structures of CBs discussed in this paper are given in Table 2.

CB-90 (2,2',3,4',5) was found to interfere with the determination of CB-101, but its contribution to the peak was <10%. It was also shown that with the SE-54 column used for analysis, CB-153 was well separated from CBs-132 and -105 (Duinker, Schulz & Petrick, 1988b).

The congeners analyzed here contribute about 60% to the total peak area of all CBs in any of the tissues (Fig. 1). As the analysis was restricted to unambiguously determined compounds (with a small, up to 10% contribution accepted for one congener, i.e., 90/101), quantitative results and comparisons can be considered reliable, as far as chromatographic overlap is concerned.

The lipid content of a tissue or organ appears to be an essential factor to determine its overall content of lipophilic compounds on a wet or dry weight basis. Therefore, we report concentrations on an hexane-extractable lipid basis. Wet weight concentrations are inadequate to establish comparisons between different organs of the same individual and between different individuals in a population or different species, because of variations in lipid contents. If required, however, they can be readily obtained from lipid based concentrations and the appropriate lipid contents given for each tissue and organ in Table 3.

The weight concentrations on a lipid basis (e.g.,  $\mu\text{g g}^{-1}$  lipid) of 7 pesticides and 17 individual chlorobiphenyls (CBs) in 87 tissues of 19 animals, which are the basic results, have been transformed into molar concentrations. Absolute concentrations are expressed as nmol (or pmol) per gram lipid. They are useful for comparing concentration levels and expressing biomagnification factors.

#### Representation of the compositions of CB- and pesticide mixtures

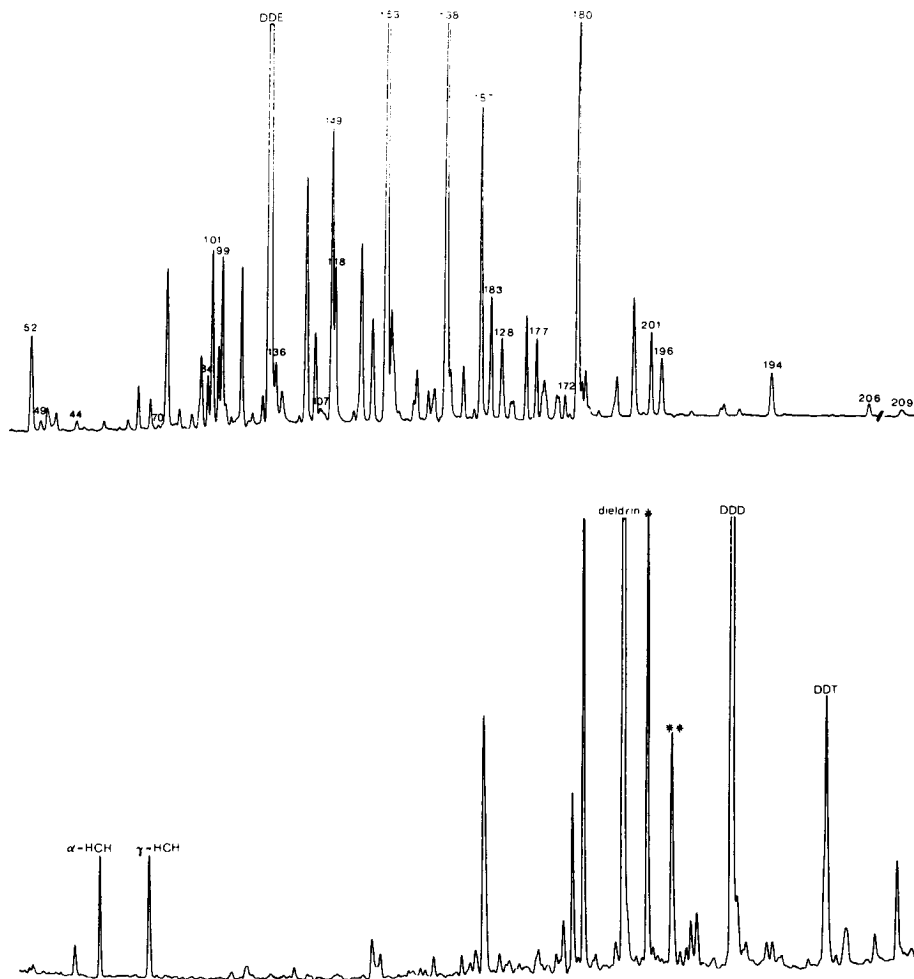
Different compositions of mixtures may occur in different tissues as result of differences in kinetics and/or mechanisms of distribution. These effects will be most easily observed from a study of the relative concentration values of the contaminant mixtures,

**Table 2.** Structure and molecular weights (M) of pesticides and chlorinated biphenyls discussed in this paper. The order in which the CBs appear in the list is the same as in the Figures and Tables (reflecting the elution pattern from a SE-54 column)

	M	Structure
<b>Pesticide</b>		
$\alpha$ -HCH	288	$\alpha$ -isomer of hexachlorocyclohexane
$\gamma$ -HCH	288	$\gamma$ -isomer of hexachlorocyclohexane
dieldrin	378	3,4,5,6,9,9-hexachloro-1a, 2, 2a, 3, 6, 6a, 7, 7a octahydro-2.7:3.6 dimethano naphth-[2,3-b] oxirene dimethanonaphthalene
HCB	282	hexachlorobenzene
p,p'-DDD	318	2,2-bis(p-chlorobiphenyl)-1, 1-dichloroethane
p,p'-DDT	352	2,2-bis(p-chlorobiphenyl)-1, 1-trichloroethane
p,p'-DDE	316	2,2-bis(p-chlorobiphenyl)-1, 1-dichloroethene
<b>Chlorinated biphenyls</b>		
18	258	2,2',5
26	258	2,3',5
52	292	2,2',5,5'
49	292	2,2',4,5'
44	292	2,2',3,5'
70	292	2,3',4',5
84	326	2,2',3,3',6
101	326	2,2',4,5,5'
99	326	2,2',4,4',5
136	361	2,2',3,3',6,6'
107	326	2,3,3',4',5
149	361	2,2',3,4',5',6
118	326	2,3',4,4',5
153	361	2,2',4,4',5,5'
138	361	2,2',3,4,4',5'
187	395	2,2',3,4',5,5',6
183	395	2,2',3,4,4',5',6
128	361	2,2',3,3',4,4'
177	395	2,2',3,3',4',5,6
172	395	2,2',3,3',4,5,5'
180	395	2,2',3,4,4',5,5'
201	430	2,2',3,3',4',5,5',6
194	430	2,2',3,3',4,4',5,5'
206	464	2,2',3,3',4,4',5,5',6
209	498	2,2',3,3',4,4',5,5',6,6'

irrespective of actual concentration levels (Duinker *et al.*, 1988). The relative concentrations of CBs have been calculated as mol % for each of the compounds in the appropriate mixture of CBs, and similarly so for the pesticides. e.g., as mol % of the total concentration. The effect of differences in absolute concentrations between samples is eliminated in this way.

The mol % contribution  $F_k^{k=1-17}(\%)$  of a chlorinated biphenyl congener k, (CB<sub>k</sub>, where k=1 ... 17) in



**Figure 1.** ECD-chromatograms of the PCB- (i.e., the first silica gel) fraction of heart of animal 16 (top) and the pesticide- (second silica) fraction of muscle of animal 6 (animals are described in Table 1).

Top: CB congeners have been identified by numbers (Table 2) according to Ballschmiter and Zell (1980). DDE = p,p'-DDE. Non-identified congeners were not analysed because of overlap problems (Duinker, Schulz & Petrick, 1988a). The strong peak of hexachlorobenzene (HCB) and the extremely weak peaks of CB-18 and CB-26 appear early in the chromatogram (not depicted here).

Bottom: Pesticides have been identified at the apex of the peaks (DDD = p,p'-DDD, DDT = p,p'-DDT). The positions where endrin and o,p'-DDD elute have been indicated by \* and \*\*, respectively. For chromatographic conditions see Duinker & Hillebrand, 1983b.

tissue  $i$  ( $t_i$  where  $i=1 \dots 6$  representing blubber, brain, heart, kidney, liver and muscle) of animal  $j$  ( $a_j$  where  $j=1 \dots 19$ ) was calculated from

$$F_k^{t_i, a_j} \% = \frac{(CB_k)^{t_i, a_j}}{\sum_{r=1}^{17} (CB_r)^{t_i, a_j}} \times 100$$

Parentheses are used to indicate molar concentrations. Calculations for the pesticides were carried

out analogously. Average mol % contributions were calculated for each tissue type (e.g., blubber) in the two groups of animals.

## Results

### *Distinction of groups of animals and groups of CB congeners*

We have treated the harbour porpoises (animals 1–11) and the other animals (12–19) as two separate

**Table 3.** Hexane extractable lipid contents (% of wet weight) in the various tissues. Averages ( $\bar{x}$ ), number (n) of data included in the statistics, relative standard deviation (s %) and range of values are given in the bottom rows. Data for *Macrocephalus* (animal 15) not included in statistics. —: tissue not analysed

Animal number	Lipid percentage in						
	Blubber	Brain	Heart	Kidney	Liver	Muscle	Sperm
1	92.9	8.2	1.5	1.8	2.3	0.5	—
2	82.0	7.5	1.7	2.5	4.1	7.9	—
3	77.8	11.2	—	1.7	8.6	0.8	—
4	67.7	5.3	—	1.7	2.3	0.7	—
5	90.6	6.5	—	—	—	2.7	—
6	93.5	10.6	2.1	6.7	2.1	0.8	—
7	88.2	13.9	1.8	2.3	3.3	1.0	—
8	—	—	—	—	3.9	—	—
9	90.2	9.4	—	2.7	2.4	—	—
10	54.2	7.2	—	1.4	6.0	1.5	—
11	72.4	6.6	—	1.9	3.2	1.0	—
12	91.7	9.5	—	—	—	1.2	—
13	79.8	8.1	—	—	—	1.1	—
14	97.9	8.3	—	—	—	0.8	—
15	92.1	18.8	18.9	—	—	18.5	97.0
16	93.9	8.7	3.5	6.0	2.5	1.2	—
17	76.2	8.3	3.5	2.1	6.7	1.5	—
18	90.8	7.4	3.2	5.0	9.6	4.1	—
19	96.1	13.4	—	—	—	2.0	—
$\bar{x}$	84.5	8.8	2.5	3.0	4.4	1.8	—
range	54.2–97.9	5.3–13.9	1.5–3.5	1.4–6.7	2.1–9.6	0.5–7.9	—
s (%)	13.5	25.6	33.0	59.4	55.2	100.6	—
n	17	17	7	12	13	16	1

groups. Harbour porpoises (group A) dwell in the coastal regions of the North Sea and represent a single species, while all other species (group B) live in the open North Sea or the Atlantic Ocean. Group B is rather heterogeneous; for each species either one, two or three animals were available. This did not justify a further subdivision of this group. The main argument to consider these animals, tentatively, as a group is the similarity of the composition of the pesticide mixtures in all tissues of these animals; their compositions differ considerably from those in the porpoise tissues (Fig. 2). Statistical analyses of the results, to be discussed below, justify this procedure for the pesticides. Differences in relative concentrations of some congeners between different species within group B could also be observed within group B (Fig. 3).

The CB congeners are divided into groups I and II, involving the congeners that were apparently metabolized (group I) and those that were persistent in harbour seals (group II, Boon *et al.*, 1987).

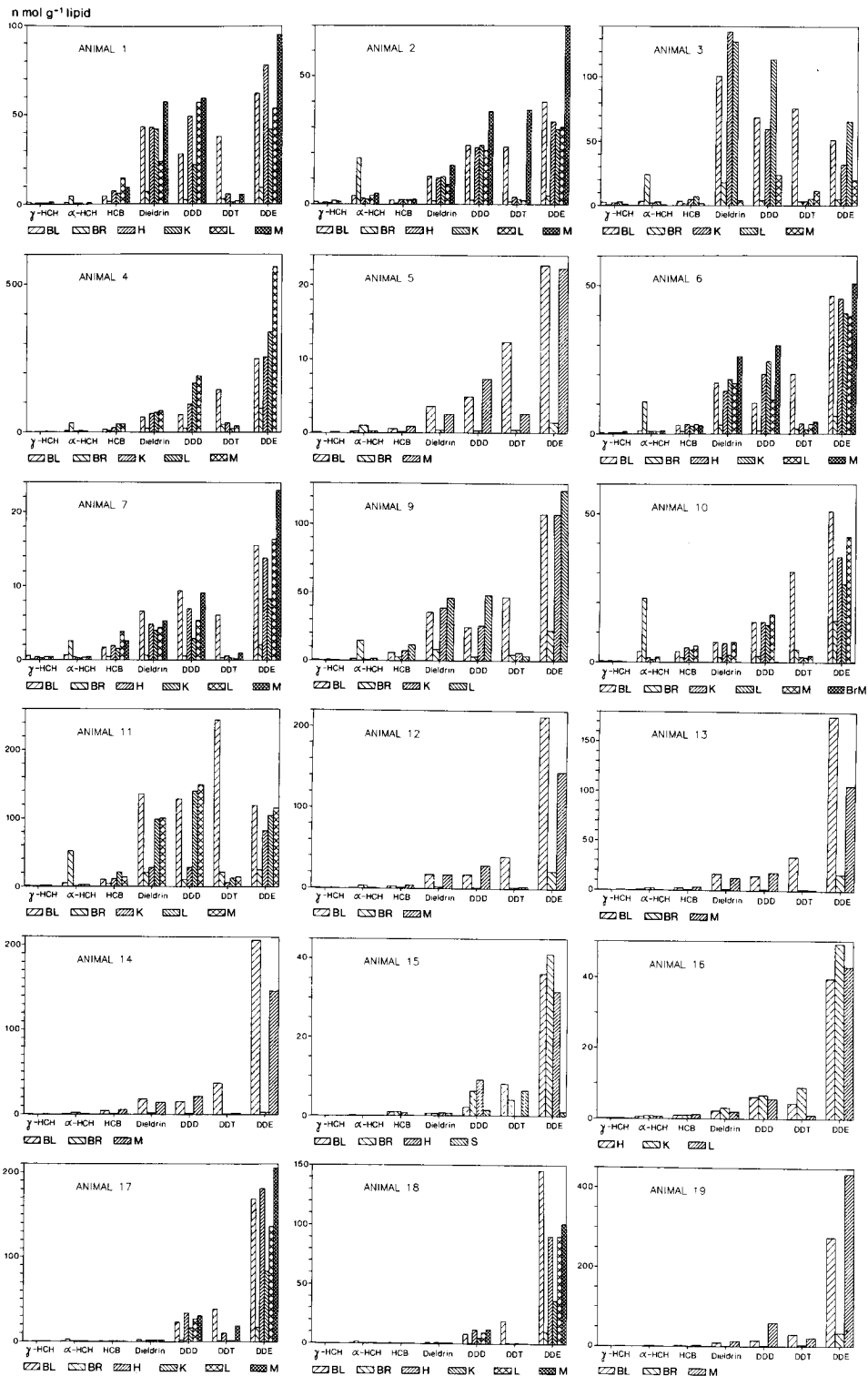
Most animals were females. Because of the limited number of animals investigated, differences in organochlorine concentrations due to differences in sex (Addison & Brodie, 1977; Tanabe *et al.*, 1981a,b,

1982; Duinker & Hillebrand, 1979b; Alzieu, Duguay & Babin, 1982) and age (Norstrom *et al.*, 1988; Muir *et al.*, 1988; Reijnders, 1979; Addison & Smith, 1974) were not recognized.

#### Lipid based concentrations

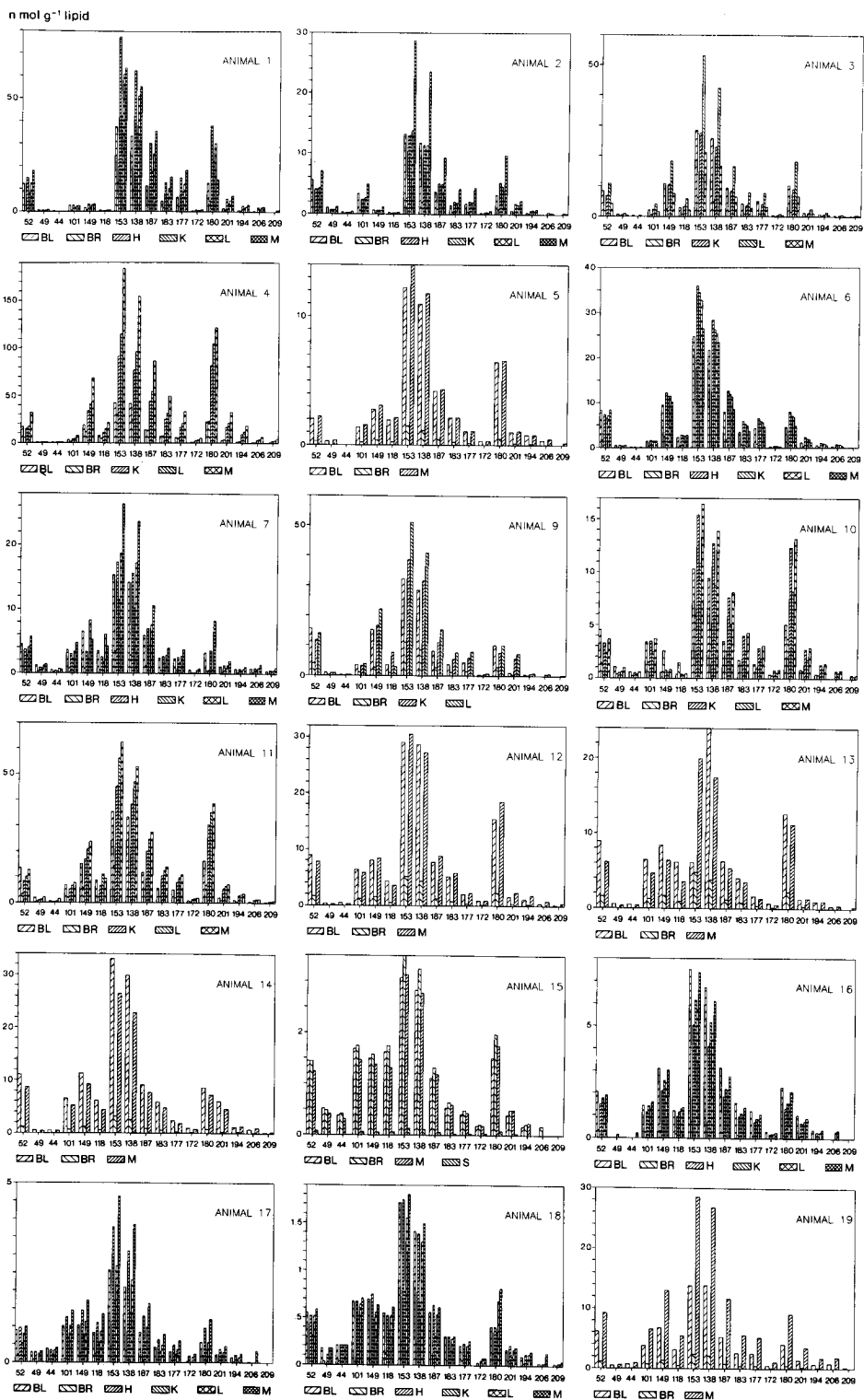
The mean lipid based concentrations of CBs and pesticides in harbour porpoises (group A) and the combined 'open sea species' (group B) are given in Tables 4 and 5 together with the matching standard deviations and the quotient of the mean concentrations of groups A and B for each compound. For statistical comparisons of the means of animal groups A and B, a Student's *t*-test was used when the samples were not significantly inhomogeneous (i.e.,  $p > 0.05$ ), as determined with an  $F_{\max}$  test with seven degrees of freedom. Often the variances of the samples were heteroscedastic, however, in which case an approximate *t*-test was used (Sokal and Rohlf, 1969).

Averages are depicted in Figures 4 and 5, including graphs for the tissues of one individual animal for each group for comparison with average values. It shows to what extent the data of an arbitrarily selected individual animal reproduces the average values for both groups A and B.



**Figure 2.** Concentrations (nmol g<sup>-1</sup> hexane extractable lipid) of  $\gamma$ -HCH,  $\alpha$ -HCH, HCB, Dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) for each tissue analysed in all individual animals 1–19. Animals 1–11 are porpoises (group A) and animals 12–19 represent other species (group B), see Table 1. The y axes have different scales. The data for animal 8 (one tissue analysed) appear in other Figures.

BL = blubber, BR = brain, K = kidney, L = liver, M = muscle, S = sperm (animal 15).



**Figure 3.** Concentrations (nmol g<sup>-1</sup> hexane extractable lipid) of CBs for each tissue analysed in all individual animals 1–19. Animal and tissue identifications as in Figures 1 and 2. The y axes have different scales. The data for animal 8 (one tissue analysed) appear in other Figures.



Most data in the literature have been presented in terms of w/w (e.g.,  $\text{ng g}^{-1}$  lipid or wet weight). Also, most concentrations have been given in terms of commercial mixture equivalents, such as Aroclor 1260. In order to allow—admittedly crude—comparisons with such literature data, concentrations of individual CBs ( $\text{ng g}^{-1}$  extractable lipid) have been totalled for each tissue and presented in Table 6.

#### *Chlorinated biphenyls*

##### *Comparison of concentrations between groups A and B*

The average lipid based concentrations of each CB congener are higher in the harbour porpoises (group A, Table 4) than in the 'open sea species' (group B, Table 4), but particularly as a result of the large standard deviations within both groups these differences are not always significant (i.e.  $p$  not always  $< 0.05$ ). The ratios of the average absolute concentrations in group A/group B are given in Table 4, together with the matching significance levels. For statistical comparisons between the same tissue type of the animal groups A (harbour porpoises) and B (combined open sea species) an  $F_{\max}$  test was performed prior to significance testing to establish whether the data to be compared had homogeneous variances. If  $s^2_{\max}/s^2_{\min} > F_{.05[2,7]}$ , a normal Student's  $t$ -test was used; else an adapted  $t$ -test for comparison of two heteroscedastic groups of samples was used (Sokal and Rohlf, 1969). The largest differences in concentrations between groups A and B are present in heart, kidney and liver, while the concentrations of only a few CB-congeners are significantly different in blubber, brain and muscle. CB-44 never differed significantly between both groups of cetaceans.

##### *Concentration ratios within groups A and B*

The ratios of lipid based concentrations of any compound X in different tissues of the same animal can be used to compare relative lipid affinities of X. For each compound, blubber/brain, blubber/heart, blubber/kidney, blubber/liver and blubber/muscle ratios have been calculated. Average ratios  $\pm$  s.dev. for each tissue of groups A and B are presented in Tables 7 and 8. Average ratios are depicted in Figures 6 and 7, including ratios for an individual animal from each group for comparison with the average values. The figures reflect the fact that in each animal, concentrations of all CBs are similar in heart, kidney, liver and muscle. They are considerably lower in brain, however, on average 4 times lower in harbour porpoises and 6–7 times lower in group B animals.

The concentration of each CB in blubber differs from those in heart, kidney, liver and muscle. It is *interesting in this respect that in harbour porpoise samples the average concentration ratio blubber/X* (X representing heart (h), kidney (k), liver (l), and

muscle (m)), decreases from around 1.0 (for the tetrachloro-CB 52) to values around 0.2–0.4 (for nonachloro-CB 206) (Table 7 and Fig. 6). In the group B species, average bl/br ratios are much higher (4–8). The bl/X ratios for the other tissues in these animals are around 1, but distinctly different among them (bl/h 0.9, bl/l 1.1, bl/m 0.9, bl/k 1.5). In these cases, the exceptional values of CB-209 have been deleted. In contrast to the situation in the harbour porpoises, they remain constant over the range of congeners (Table 7, Fig. 6). These differences between blubber and the other tissues are not obvious from visual inspection of the CB patterns. In the following, we shall consider the patterns in blubber and the other tissues as similar, realizing that there are the differences mentioned above. We feel that more data are required to establish whether the observed effects are real and may have wider validity.

#### *Pesticides*

##### *Comparison of concentrations between groups A and B*

The lipid based concentrations of the pesticides  $\alpha$ - and  $\gamma$ -HCH, HCB and dieldrin are higher in most tissues of the harbour porpoise compared to the open sea species (Table 5). In contrast, the concentrations of p,p'-DDE were higher in the open sea animals. Differences of p,p'-DDD and p,p'-DDD and p,p'-DDT are small compared to the other pesticides and are not significant.

##### *Concentration ratios within groups A and B*

The differences in concentrations in blubber relative to brain are larger for pesticides than for CBs. Average bl/br ratios increase in the series HCB (3), dieldrin and DDE (6–8), DDD (12), DDT (16). In group B animals, even higher average ratios occur for DDE (19) and DDT (24).  $\alpha$ -HCH, however, had a high concentration in brain lipids: average bl/br ratios are considerably smaller than 1.0 (about 0.2 in both groups). The large relative concentration of DDT in blubber is obvious for both groups of animals (bl/X ratios between 8 and 20 in animals of group A and between 20–60 in animals of group B). Much smaller differences in concentration ratios were found for the other compounds and tissue types (thus, excluding  $\alpha$ -HCH, DDT and brain).

#### *Compositions of CB- and pesticide mixtures*

##### *Chlorinated biphenyl patterns*

##### *Distribution between various tissues of the same animal*

The compositions of CB-mixtures (whereby each concentration is expressed as mol % of the sum of all concentrations) are much more similar between different tissues of the same animal than absolute concentrations (compare Tables 4 and 9). So, the concentrations of many CB congeners covary

**Table 4.** Average concentrations (avg) of CBs (in nmol g<sup>-1</sup> lipid) and relative standard deviation (% of average) in the various tissues of animals 1-11 (group A) and 12-19 (group B). For numbering of animals see Table 1 and for numbering of CBs see Table 2. No entry: data do not allow a reliable quantitation. Significance levels for differences between groups A and B (two-tailed *t*-test) \**p*<0.05; \*\**p*<0.01; \*\*\**p*<0.001. Group I and II congeners are separated by an empty row.

Tissue	Blubber				Brain				Heart			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
CBs												
52	9.2	53	5.0	79	2.2	69	0.9	64	7.4	60	1.0	39
49	0.8	49	0.4	51	0.1	98	0.1	213	0.7	22	0.1	141
44	0.3	58	0.4	50	0.1	137	0.1	207	0.2	68	0.2	77
101	3.1	46	3.5	70	0.9	60	0.7	77	2.5	21	1.0	24
149	8.2	78	5.1	73	2.9	111	1.0	61	4.8	91	1.4	38
118	3.2	88	3.0	72	1.1	117	0.6	83	1.4	87	0.9	28
153	25.2	44	12.1	95	9.4	85	2.5	70	36.0	71	3.5	38
138	23.1	46	13.7	83	7.4	87	2.1	70	29.5	68	2.9	39
187	8.0	43	4.2	73	3.4	105	0.7	64	13.8	72	1.3	39
183	3.6	48	2.5	81	1.6	122	0.4	66	5.8	73	0.6	42
177	3.7	49	1.3	65	1.4	95	0.2	62	6.5	79	0.5	41
172	0.4	37	0.4	80	0.1	142	0.1	74	0.3	56	0.1	70
180	9.4	63	5.7	95	4.5	143	1.0	89	12.9	114	0.9	42
201	1.3	41	1.5	118	0.7	125	0.2	68	2.7	62	0.4	46
194	0.7	41	0.6	71	0.3	124	0.1	96	1.2	66	0.2	33
206	0.3	82	0.3	114	0.1	193	0.0	170	0.9	74	0.0	
209	0.1	154	0.0	138	0.1	192	0.0	185	0.1	173	0.0	
	Kidney				Liver				Muscle			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
CBs												
52	7.3	51	0.9	74	8.8	51	0.9	44	10.4	85	4.6	76
49	0.6	34	0.1	72	0.7	36	0.1	73	0.9	54	0.3	55
44	0.2	79	0.1	70	0.3	72	0.2	77	0.4	101	0.4	54
101	2.9	38	0.9	44	3.4	39	1.0	26	4.0	58	3.4	64
149	10.5	98	1.3	73	13.8	85	1.4	57	13.7	148	5.5	76
118	3.7	96	0.7	49	5.0	92	0.9	35	4.8	133	2.7	61
153	35.5	65	2.9	81	45.1	64	3.3	52	49.2	102	15.3	75
138	29.4	66	2.4	82	37.4	64	2.8	51	41.7	102	13.6	77
187	14.4	80	1.0	82	18.3	76	1.2	55	21.8	114	4.9	76
183	7.2	95	0.5	79	8.9	92	0.6	53	11.0	129	2.8	77
177	6.1	71	0.4	83	7.6	70	0.5	56	9.1	108	1.6	93
172	0.7	99	0.1	102	0.9	98	0.1	36	1.2	122	0.5	70
180	19.9	118	0.8	75	23.3	125	0.9	48	24.8	143	6.5	90
201	4.4	104	0.3	77	5.3	105	0.4	63	6.3	150	1.7	88
194	2.0	127	0.1	57	2.5	129	0.2	40	3.3	160	0.8	76
206	0.9	82	0.0	141	1.3	83	0.1	127	1.4	123	0.6	91
209	0.4	135	0.0	141	0.5	116	0.0	141	0.8	152	0.1	130

between organs and tissues of the same animal (Fig. 8). Also, a large similarity among all porpoises and among all open sea animals is observed (Fig. 8;

Table 9). The differences between blubber and the other tissues was mentioned in the section titled 'Chlorinated biphenyls'.

Table 4. Continued

CBs	Ratio of averages A/B					
	Blubber	Brain	Heart	Kidney	Liver	Muscle
52	1.8	2.4*	7.6***	8.5***	9.4***	2.3
49	2.4**	1.9	7.1***	7.2***	5.4***	3.0**
44	0.8	0.8	0.9	1.5	1.5	1.1
101	0.9	1.2	2.4***	3.3***	3.5***	1.2
149	1.6	2.9	3.4*	8.4*	10.1**	2.5
118	1.1	1.9	1.7	5.3*	5.6*	1.8
153	2.1*	3.7*	10.2**	12.3***	13.6***	3.2
138	1.7	3.5*	10.3**	12.2**	13.5***	3.1
187	1.9*	4.6*	11.0**	14.3**	15.1**	4.4
183	1.4	3.8	9.2**	13.9**	14.8**	3.9
177	2.8**	5.6*	13.8**	15.5**	16.7**	5.7*
172	1.0	1.7	2.9**	8.6*	6.4*	2.4
180	1.7	4.5	14.3*	26.2*	24.9*	3.8
201	0.8	3.4	6.5*	14.5*	13.3*	3.7
194	1.1	3.6	6.1**	14.6*	13.8*	4.0
206	1.0	2.3		131.3**	11.6**	2.4
209	1.7	7.7		63.0*	76.2*	6.7

*Comparison of peak patterns between groups A and B*  
Differences exist between the groups of animals: the ratio of the mean mol % contributions to  $\sum$ CB (i.e., A/B) of the potentially metabolizable congeners of group I in the groups of animals A and B is virtually always < 1, but for the persistent congeners of group II this ratio is often > 1. However, these differences are generally significant for only a few congeners (< 1: CBs -44, -101, -118; > 1: CBs -153, -138, -187, -177). Significance testing was performed in the same way as for comparison of the absolute CB concentrations (compare 'Chlorinated biphenyls'). CB-172 behaved in an unexpected way in that it belongs to group II (persistent congeners), but its ratio A/B was also often significantly < 1 (Table 9).

In order to investigate in more detail whether not only the absolute concentrations but also the relative concentrations of chlorobiphenyls in the tissues of animals differ between groups A and B, a two-way analysis of variance was carried out. Concentrations were transformed into concentration ratios relative to CB-153, i.e., X/153. This enables the concentration patterns to be compared, taking into account both potential effects, viz., differences between the groups A and B and differences between the organs within each group. The underlying significance level was set at  $p=0.05$ .

The combined (main) effect (where  $p$  values < 0.05 indicate significant differences) indicates that significant overall differences exist for several CBs (Table 10). Comparison with the values for the effects of the two groups (A and B) and the organs,

indicates that the combined effects arise mainly because of differences between the groups A and B, and not because of differences between tissues within each group. The difference for CB-206 ( $p < \alpha$  in the overall effect) is caused by differences between the organs and not between the animal groups. It is also observed that differences in average values were not significant for several other CBs. Most of these cases correspond to non-significant differences between the groups as well as between the organs. In conclusion, significant differences in peak patterns of CBs are present between groups A and B, taking into account the organs. Another conclusion is that biotransformation is stronger in porpoises than in the open sea species.

#### *Pesticides*

##### *Distribution between various tissues of the same animal*

A large similarity is observed between the mol % contributions of the pesticides in blubber of groups A and B (Table 5). The same applies to each of the other tissues for both groups of animals. In contrast with the situation for the CBs, the patterns for the pesticides differ significantly between the tissues, however. Despite considerable concentration differences between tissues of the harbour porpoises (Table 5), a reasonable similarity is observed for the mol % contributions of  $\gamma$ -HCH, dieldrin and DDE (Fig. 9). Differences are larger for  $\alpha$ -HCH, HCB, DDD and DDT. Contributions of DDD in blubber are smaller than in liver but those of DDT are higher. The large

**Table 5.** Average (avg) concentrations (nmol g<sup>-1</sup> hexane extractable lipid) ± % standard deviation (std%) for each of the pesticides in different tissue types of the porpoises (group A) and of the other species (group B). For numbering of animals see Table 1. Significance levels for differences between groups A and B (two-tailed t-test) \*p < 0.05; \*\*p < 0.01; \*\*\*p < 0.001.

Animal group	Blubber				Brain				Heart			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
gamma-HCH	0.8	74	0.2	25	0.1	72	0.0	245	0.5	24	0.1	71
alpha-HCH	2.4	73	0.4	29	17.8	81	1.8	47	1.0	76	0.4	68
HCB	4.3	71	1.8	72	1.8	86	0.6	60	3.6	62	0.6	34
Dieldrin	41.1	103	9.3	81	7.2	96	1.0	65	18.2	81	1.4	49
DDD	36.9	99	13.5	46	3.8	106	1.9	98	24.6	62	15.1	72
DDT	63.9	111	29.3	37	6.5	133	1.8	64	3.1	57	3.6	109
DDE	76.4	86	174.5	39	16.9	135	20.3	60	42.4	55	85.9	69

Animal group	Kidney				Liver				Muscle			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
gamma-HCH	0.6	89	0.1	13	0.8	103	0.1	27	0.7	62	0.2	28
alpha-HCH	1.1	62	0.4	58	1.6	83	0.4	32	1.6	91	0.4	17
HCB	5.8	67	0.5	64	9.9	84	0.7	54	6.6	123	2.6	67
Dieldrin	38.5	101	1.7	62	40.1	105	1.3	43	28.9	115	10.1	63
DDD	32.5	80	9.3	56	58.7	97	13.8	66	35.2	117	28.2	54
DDT	5.8	158	3.6	105	4.0	100	0.7	46	9.9	112	7.3	117
DDE	70.1	101	56.4	36	83.3	109	89.7	43	99.6	157	189.6	61

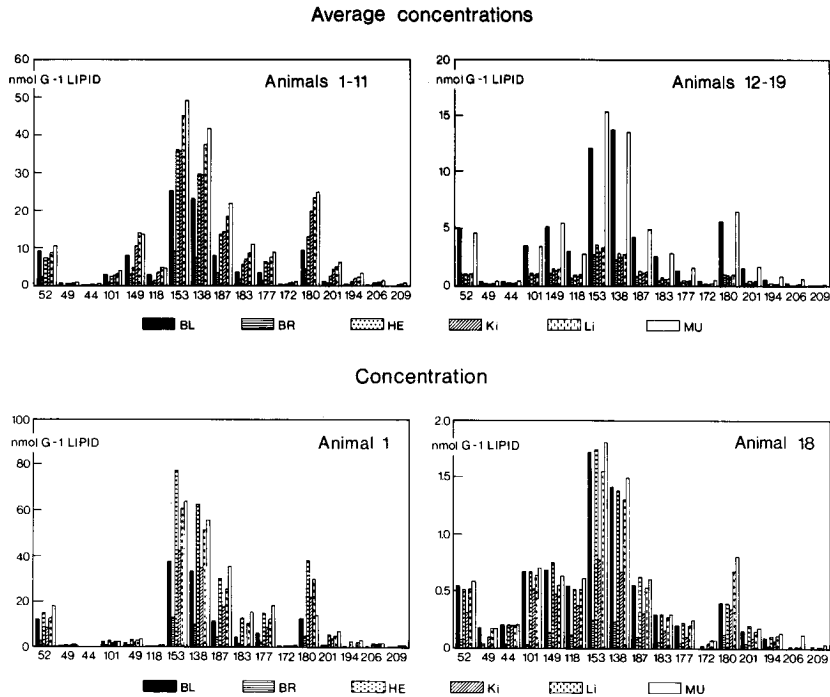
	Ratio of averages A/B					
	Blubber	Brain	Heart	Kidney	Liver	Muscle
gamma-HCH	5.0**	27.3**	5.0***	4.6*	7.5*	4.4**
alpha-HCH	6.2**	9.8**	2.8*	2.6*	4.2*	4.1*
HCB	2.4*	3.0*	5.8**	12.1**	15.2**	2.5
Dieldrin	4.4*	7.1*	12.9**	22.8*	30.5*	2.9
DDD	2.7	2.0	1.6	3.5*	4.3*	1.2
DDT	2.2	3.6	0.9	1.6	5.9*	1.4
DDE	0.4**	0.8	0.5	1.2	0.9	0.5

mol % contributions of  $\alpha$ -HCH in brain, of DDT in blubber and DDE in all tissues are striking features.

The relative concentrations of the 'DDTs' in porpoises are DDD < DDT < DDE in blubber and brain and DDT < DDD < DDE in the other tissues (Fig. 9). These patterns are different in the open sea species (group B): considerably smaller mol % contributions occur for  $\alpha$ -HCH in brain and very large (up to 80 mol % contributions) for DDE in all tissues (Fig. 9). The relative concentrations of DDD and DDT in blubber and liver are similar to those in the porpoises. In group B animals the order is the same as in group A but relative concentrations

are different: DDD < DDT < DDE in blubber and brain and DDT < DDD < DDE in the other tissues.

*Comparison of peak patterns between groups A and B*  
The same statistical approach (two-way analysis of variance) applied to the concentration ratios of CBs in different tissues of both groups A and B was also applied to the pesticides, in order to detect any significant differences between the groups A and B, taking into account the various organs. For all pesticides, significant differences in the concentration ratios X/DDE were found (Table 10). As for the CBs, the differences result mainly from differences



**Figure 4.** Concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of CB congeners (identified by numbers, Table 2) in blubber (BL), brain (BR), heart (HE), kidney (KI), liver (LI) and muscle (MU) of animals 1–11 (group A, left) and animals 12–19 (group B, right). The y-axes have different scales. Top: average values for groups A and B. Bottom: values for individual animals 1 and 18.

between the groups A and B and not from differences between the organs.

## Discussion

### *Chlorobiphenyls*

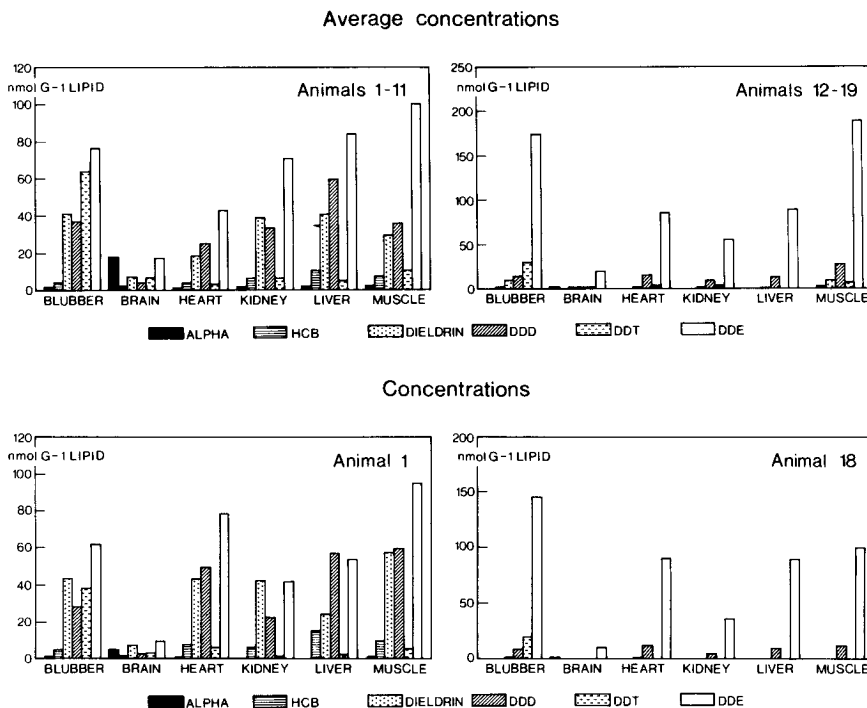
#### *Chlorobiphenyl patterns in relation to metabolism*

In marine mammals, food is the dominant source of contaminants, direct uptake from water being insignificant because marine mammals do not drink seawater. Blood probably plays an important role as transport compartment between different tissues (Harding & Addison, 1986). It is recognized that several congeners (hexa- and heptachlorocongeners, etc.) are persistent in the environment. CB-153 is a well known example. As it occurs with the highest concentrations of all CBs in mammal tissues, the X/153 concentration ratio is a convenient measure to estimate the effect of metabolism on the fate of X. This facilitates the comparison with the data of Boon *et al.* (1987). These authors studied the kinetics of CBs in female harbour seals. One group of animals (originating from the Dutch Wadden Sea and the Wash) was fed with Atlantic fish (mackerel) and the other group with Wadden Sea fish (plaice).

CB patterns were studied in food, faeces and blood as a function of time. Highly similar patterns were found for each compartment, but patterns differed among them. Since no differences were found in CB patterns between different tissues in the present study, we assume that the CB pattern in seal blood represented the pattern in the whole seal.

Assuming the necessity of the formation of an arene oxide in the biotransformation of CBs (Safe 1984; Safe *et al.*, 1985), the differences in CB patterns between seal blood and fish most likely resulted from enzymatic bio-transformation of CBs possessing vicinal H-atoms in the m,p position or in the o,m position in combination with not more than one ortho-chlorine. In the following we make use of the data in blood and faeces of seals measured during the experiments of Boon *et al.* (1987) and of the data in tissues of cetaceans reported here. We used the original chromatograms of seal blood to calculate data on CBs which were not reported by the latter authors. As they did not measure CB-206, we have deleted this congener from the Figures and Tables that will follow.

The graphs representing X/153 concentration ratios in the paper of Boon *et al.* (1987) result in the following observations:



**Figure 5.** Concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of  $\alpha$ -HCH (ALPHA), HCB, dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) in blubber, brain, heart, kidney, liver and muscle of animals 1-11 (group A, left) and animals 12-19 (group B, right). The small values of  $\gamma$ -HCH have not been included. The y-axes have different scales. Top: average values for groups A and B. Bottom: values for individual animals 1 and 18.

(i) the CB patterns in Wadden Sea fish (plaice) and Atlantic fish (mackerel) are very similar, although concentration levels differ by an order of magnitude.  
 (ii) the same CB pattern is found in faeces of animals fed with Wadden Sea or Atlantic fish.  
 (iii) the CB pattern in seal blood remained unchanged after feeding with mackerel for two years while the concentrations decreased by 50% in the same period.

According to the model of Boon *et al.* (1987) the following CB congeners reported here, should be metabolizable in seals: 18, 26, 44, 49, 52, 70, 84, 92, 101, 107, 118, 136 and 149. According to the model presented by Tanabe *et al.* (1988), only congeners with vicinal H-atoms in the o,m position with less than two ortho chlorines are metabolizable in cetaceans. This excludes all congeners except CBs -26, -107 and -118 from the list mentioned above for seals. CBs -99, -128, -138, -153, -172, -177, -180, -183, -187, -194, -201, -206 and -209 are persistent to enzymatic metabolism in both models.

Figure 10 shows the PCB patterns normalized to CB-153 in fish from the Dutch Wadden Sea, blood of seals fed with this fish, harbour porpoise

blubber and blubber of the open sea cetaceans. In the following, seal blood and cetacean blubber are assumed to represent the situation in the entire animal. According to Table 9, this appears to be legitimate in cetaceans. The relative concentrations of the major persistent congeners resemble each other closely or are higher in mammals than in fish. Decreased relative concentrations are observed for CB-172 in porpoise blubber and CB-209 in both groups of cetaceans. When the relative concentration to CB-153 of a given congener is considerably lower in a marine mammal tissue than in fish, this indicates metabolism of that congener. This is observed for all potentially metabolizable congeners in seals, but not for CBs-52, -136, and -149 in both groups of cetaceans and CBs -84 and -101 in the cetaceans from the open sea. So, the order of capacity for biotransformation decreases in the order seals > harbour porpoises > open sea cetaceans > fish.

In the cetaceans, the CBs present in lowered relative concentrations compared to fish include all congeners according to the model of Tanabe *et al.* (1988), but also CBs -44 and -49 in both groups of cetaceans and CBs -101 and -84 in the harbour

**Table 6.** Sum of concentrations of 17 individual CBs discussed in this paper in tissues of porpoises (animals 1-11 group A) and the other species (animals 12-19 group B). For numbering of animals see Table 1 and of CBs see Table 2. Concentrations in  $\mu\text{g}^{-1}$  lipid. No entry: tissue not analysed.

Animal number	Blubber	Brain	Heart	Kidney	Liver	Muscle
1	126	40	269	151	219	241
2	47	9	49	48	49	99
3	110	22		101	191	74
4	186	132		434	543	825
5	48	5				52
6	91	20	126	118	114	97
7	67	12	63	45	77	104
8					155	
9	128	38		148	187	
10	43	28		70	48	75
11	158	58		201	248	277
12	120	20				125
13	89	19				84
14	134	12				107
15	17	20				17
16	32	3	21	25	24	31
17	12	2	16	8	12	20
18	8	5	8	5	8	9
19	66	11				131

porpoises (group A). All these congeners except CB-101 do contain vicinal H-atoms in the o,m position, but in combination with more than one ortho-Cl leading to persistence to biotransformation in the models of Boon *et al.* (1987) as well as of Tanabe *et al.* (1988). The apparent persistence of CBs -99, -128, -138 and -177 support the hypothesis that for the occurrence of biotransformation in the o,m region a maximum of one ortho-Cl is allowed. However, CBs -44, -49, -84, and -101 all contain vicinal H-atoms in the m,p positions. Therefore some biotransformation activity with respect to the m,p region of PCBs cannot be excluded for cetaceans in general and more specifically for *P. phocoena*, but it is certainly lower than for harbour seals. The highest relative concentrations of potentially metabolizable CBs are found in *Physeter macrocephalus* (animal 15, Fig. 3) and *Sotalia fluviatilis* (animals 17 and 18). These animals possessed the lowest absolute lipid based CB concentrations of all animals analysed (Table 6). This may reflect increased MFO-activity at higher CB concentrations or general physiological differences between species.

Using multidimensional gas chromatography (MDGC, Duinker, Schulz & Petrick, 1988a), we found that three toxic CBs, i.e., CB-77 (3,3',4,4'), CB-118 (2,3',4,4',5) and CB-105 (2,3,3',4,4') were present in measurable concentrations. CB-77 appeared to

contribute about 10% to the peak 110/77 and CB-105 represented about 7% of CB-153 (data for CB-118 are included in the paper). Extremely low levels of CBs can be quantitated by MDGC-ECD. However, the toxic congeners -37, -81, -123, -114, -126, -167, -156, -157, -169 and -189 were below detection limits.

This finding is likely to be of toxicological importance, since the parent compounds of the planar non-ortho chlorine containing congeners of their mono- and di-ortho analogues have the same toxicological mechanisms as 2,3,7,8 TCDD in rats, i.e., they bind to the cytosolic arylhydrocarbon receptor, show a 3-MC or mixed 3-MC and PB-type induction of MFO enzymes, cause liver hypertrophy and hepatic porphyria (Safe, 1984; Safe *et al.* 1985). However, a metabolite of CB-77 showed an entirely different kind of toxicity in mice and rats (Brouwer *et al.* 1986): it caused lowered plasma levels of retinol (vitamin A) and thyroxine.

Studies of persistent CBs in mammal tissues (as has been the case almost exclusively) do not reveal the presence or absence (because of metabolism) of congeners which have structural properties allowing them to be metabolized and possibly to cause harmful effects. Clearly, experimental studies (e.g., such as carried out by Reijnders (1986), involving both metabolizable and persistent congeners are necessary to obtain more insight into their distribution,

**Table 7.** Hexane extractable lipid based concentration ratios blubber/brain, blubber/heart, blubber/kidney, blubber/liver and blubber/muscle for each CB identified by numbers (Table 2). Presented are the average ratios (avg) and their relative standard deviations (% of average, std%) for the tissues in harbour porpoises (group A) and in other species (group B, Table 1). Concentrations are in nmol g<sup>-1</sup> lipid. No entry: data do not allow a reliable quantitation

Animal group	Blubber/Brain				Blubber/Heart				Blubber/Kidney			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
52	4.6	22	5.8	41	1.1	18	0.8	73	1.4	13	1.6	18
49	4.2	14	3.4	50	1.2	16			1.5	25	0.8	100
44	4.1	22	4.1	55	1.4	18	1.0	0	1.3	23	1.4	29
101	4.2	33	8.2	73	1.1	19	0.8	72	1.2	25	1.3	15
149	4.3	60	5.7	49	1.1	50	0.8	76	1.1	61	1.4	8
118	4.6	53	6.0	50	1.4	27	0.8	72	1.3	61	1.4	16
153	3.8	50	5.9	57	0.8	26	0.8	75	0.9	27	1.6	25
138	4.4	50	7.1	46	0.8	23	0.9	76	0.9	25	1.6	21
187	4.0	61	6.6	51	0.6	27	0.9	81	0.7	41	1.6	11
183	4.2	63	6.6	46	0.6	31	0.9	77	0.7	50	1.6	17
177	4.2	52	6.3	51	0.7	27	0.9	80	0.8	42	1.6	19
172	2.9	70	4.1	85	1.6	91	1.7	0	1.4	138	0.8	100
180	4.9	75	6.9	65	2.1	132	0.9	77	0.7	45	1.4	13
201	3.4	69	8.0	82	0.5	49	0.8	81	0.5	76	1.5	18
194	3.5	65	5.0	45	0.5	52	0.7	77	0.6	72	1.5	6
206	4.7	53	3.9	173	0.5	63			0.4	109	0.0	
209	1.7	112	1.2	100	0.8	0			0.3	145	0.0	

Animal group	Blubber/Liver				Blubber/Muscle			
	A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%
52	1.2	26	1.2	11	1.0	41	1.1	21
49	1.2	27	1.1	7	1.0	35	1.2	26
44	1.1	29	1.1	5	0.7	32	0.9	42
101	1.0	25	1.1	6	0.9	26	1.0	26
149	0.8	27	1.1	15	0.8	44	1.0	27
118	0.9	34	1.0	5	0.8	29	1.1	35
153	0.7	26	1.1	13	0.7	44	0.8	37
138	0.8	25	1.1	19	0.7	46	1.0	30
187	0.6	28	1.1	23	0.6	62	0.9	30
183	0.6	32	1.2	22	0.6	68	0.9	29
177	0.6	28	1.1	24	0.7	66	0.9	32
172	0.7	40	0.5	141	0.6	51	0.8	66
180	1.2	146	1.0	35	0.7	62	0.8	36
201	0.4	55	1.0	25	0.5	74	0.8	35
194	0.5	50	1.0	21	0.6	71	0.7	26
206	0.3	108	0.0		0.6	128	0.3	91
209	0.2	150						

kinetics and effects in marine mammals. Identification of metabolites of CBs is complicated, but it can be achieved (Jensen and Jansson, 1976).

#### *Biomagnification of chlorobiphenyls*

Figure 11 (left) represents average lipid based concentrations of all CBs listed in Table 2 for

Wadden Sea fish, blood of seals fed with Wadden Sea fish (i.e., the Wadden Sea group) for two years, and of porpoise blubber. Concentrations of persistent congeners (group II) in blood of seals of the Wadden Sea group and blubber of porpoises (group A animals) are larger than those in Wadden Sea fish (roughly a factor of 2–4) but those of most of



**Table 8.** Concentration ratios blubber/brain, blubber/heart, blubber/kidney, blubber/liver and blubber/muscle for each of the pesticides. Presented are average ratios (avg) and their relative standard deviations (% of average, std%) for the tissues in the porpoises (group A) and the other species (group B). Concentrations are in nmol g<sup>-1</sup> hexane extractable lipid. For numbering of animals, see Table 1

Animal group	Blubber/Brain				Blubber/Heart				Blubber/Kidney			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
gamma-HCH	6.0	25	4.0	0	1.4	18	1.2	14	2.8	143	1.4	6
alpha-HCH	0.2	34	0.2	15	1.6	16	1.1	6	2.9	85	1.6	8
HCB	2.9	24	3.3	36	0.8	14	0.9	27	0.9	18	1.2	2
Dieldrin	6.6	34	7.8	45	1.2	12	1.0	9	1.4	85	1.5	1
DDD	11.9	35	12.4	46	0.9	40	0.5	40	1.6	83	1.6	10
DDT	16.0	50	23.5	50	7.8	23	27.6	86	19.8	57	31.2	3
DDE	7.7	51	19.4	132	1.0	16	1.2	23	1.4	20	3.0	33

Animal group	Blubber/Liver				Blubber/Muscle			
	A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%
gamma-HCH	1.5	43	1.8	14	1.2	64	1.1	13
alpha-HCH	1.7	59	1.4	6	1.7	65	1.1	12
HCB	0.6	40	0.8	1	0.8	59	0.7	4
Dieldrin	1.4	47	1.5	3	3.8	210	1.1	19
DDD	0.9	48	0.9	1	0.9	83	0.6	29
DDT	18.5	52	60.3	11	7.5	61	20.4	74
DDE	1.1	31	1.4	13	1.0	60	1.2	30

the metabolizable congeners (group I) are similar to or smaller than those in fish. Figure 11 (right) shows similar data for Atlantic fish, blood of seals fed with Atlantic fish (i.e., the Atlantic group) for two years and blubber of group B animals. Concentrations in blood of seals of the Atlantic group and blubber of group B animals are one to two orders of magnitude higher than those in Atlantic fish.

Thus a stronger biomagnification is observed in the seals fed with Atlantic mackerel and in the open sea cetaceans than in the seals fed with Wadden Sea fish and the harbour porpoises. However, this conclusion is based on only three chromatograms of Atlantic fish and should be further investigated. Also, the seals fed with Atlantic fish showed decreasing concentrations in blood and may not have reached their equilibrium level with the relatively daily intake of contaminants from oceanic fish. The influence of different rates of metabolism on the degree of biomagnification of specific CB congeners in the different mammals is clearly observed.

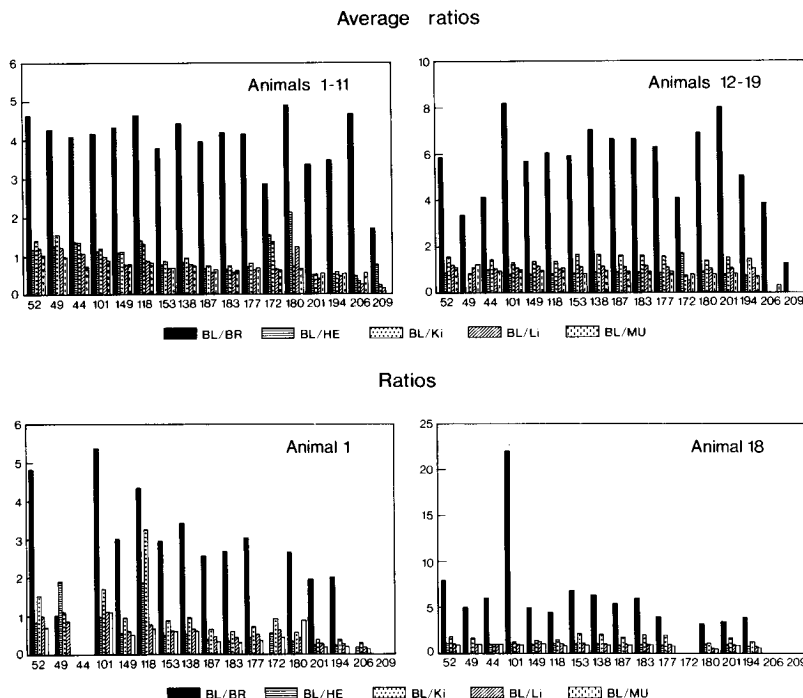
#### *Distribution of CBs between tissues*

The results discussed above indicate a clear pro-

portionality between the concentration of each CB between different compartments.

Figures 3 and 8 and Table 9 show that the composition of the CB mixtures in blubber, brain, heart, kidney, liver and muscle of each animal investigated are virtually constant. Also, differences between animals of the same species from the same area are only minor, irrespective of absolute concentrations. We have found no indication that the patterns in brain are different from those in other compartments. Tanabe *et al.* (1988), however, reported a larger contribution of early eluting congeners in brains of striped dolphins (it should be noted that these congeners were not specified, but their degree of chlorination can be derived from the Figures).

The present results show that OCs do not accumulate in blubber at the expense of levels in other tissues. Such behaviour can be described by kinetic models, e.g., the typical mammillary model applied in pharmacology and toxicology (Robinson & Roberts 1968; Robinson, Baldwin & Walker, 1969; Moriarty, 1975, 1978). Here, blood plays an important role of the central transport compartment acting as a bridge between the outside world and a series of



**Figure 6.** Ratios of concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of CB congeners (identified by numbers, Table 2) between blubber and other tissues (blubber/brain (BL/BR), blubber/heart (BL/HE), blubber/kidney (BL/KI), blubber/liver (BL/LI), and blubber/muscle (BL/MU)) in animals 1–11 (group A, left) and animals 12–19 (group B, right). Top: average values for groups A and B. Bottom: values for individual animals 1 and 18. The y-axes have different scales.

peripheral compartments (organs) which are more or less independent from each other while maintaining a dynamic balance with the blood and among them (Matthews & Dedrick, 1984; Aguilar, 1985).

A retention kinetics model such as described for the preferential accumulation of Pb in bone tissue (Morgan & Roan, 1972) does not correctly describe the kinetics of organochlorines in fatty tissues of mammals.

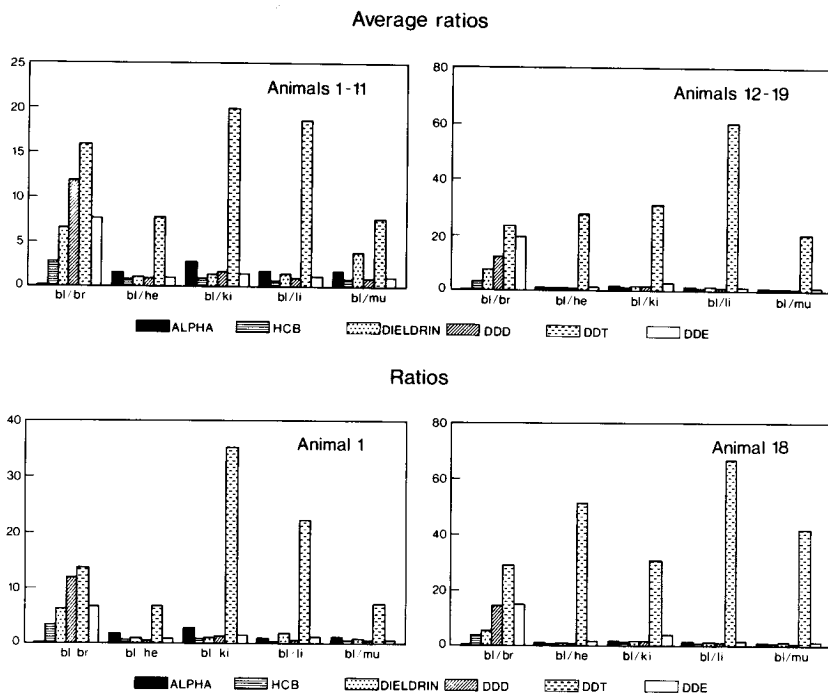
Uptake of CBs occurs mainly through food, followed by partitioning into blood cells, haemolymph and blood plasma lipo-proteins and subsequently into the non-polar moieties of different tissues (Matthews & Dedrick, 1984; Harding & Addison, 1986). There are several possibilities for deviations in the correlations between levels in blubber and other compartments (Aguilar, 1985). According to Moriarty (1983) such relations often break down in field populations, essentially because the condition of (near-) steady state is not fulfilled, but our data support this model to a large extent.

By far the most studies of organochlorines in marine mammal tissues are related to their concen-

trations in blubber. Other tissues have been studied in relatively few cases (Reuthergaard & Knap, 1987). Little is known about the uptake of organochlorines by marine mammals; even less is known about the mechanisms of transport and distribution within an animal. Detailed data involving several tissues are required for developing kinetic transport models in marine mammals. For instance, there is no information for marine mammals on what happens with organochlorines when depot lipids are mobilized for energy purposes. As blubber is a complex and inhomogeneous tissue of variable volume, various problems arise in the interpretation of blubber data only (Aguilar 1985).

We shall analyze here to what extent data on individual congeners in different tissues of various individuals and species may help to understand the distribution mechanisms of lipophilic compounds between tissues.

In order to detect differences in absolute concentrations of each congener between the organs within each of the groups A and B, a one-way analysis of variance was carried out ( $\alpha=0.05$ ). The results are



**Figure 7.** Ratios of concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of  $\alpha$ -HCH (ALPHA), HCB, dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) between blubber and the other tissues (blubber/brain (BL/BR), blubber/heart (BL/HE), blubber/kidney (BL/KI), blubber/liver (BL/LI) and blubber/muscle (BL/MU)) in animals 1–11 (group A, left) and animals 12–19 (group B, right). The y-axes have different scales. Top: average values for groups A and B. Bottom: values for individual animals 1 and 18.

summarized in Table 11. Significant differences are identified in the columns 'ONE WAY' for both groups of animals. Since this analysis identifies whether there is a difference or not, without identifying the cause of any difference, a Tukey test was also performed: the columns 'Tukey test' identify the organs, where the average concentrations differed significantly from others for each CB congener. There are no differences for twelve congeners for both groups, and significant differences for five congeners, which are not the same in groups A and B (except CB Nos. 49 and 44). In all cases where significant differences occur, brain tissue is involved (with the lowest concentrations) and in practically all cases also muscle (with the highest concentrations).

The patterns of persistent CBs (i.e., group II) appear to be very similar in seals and porpoises from the coastal North Sea region. The same similarity applies to the combined 'open sea species' of group B in the present investigation (Fig. 3, Table 9). Also, CB mixtures of persistent congeners analyzed in the fatty protuberance of the melon from Commerson's dolphins caught in the coastal waters

of the Kerguelen Islands (Southern Indian Ocean) appear to be very similar (Abarnou, Robineau & Michel, 1986) to the ones that we are reporting here for the North Atlantic. It should be noted that concentration levels in mammal tissues and fish from these remote areas are one to two orders of magnitude below levels reported here for the North Atlantic. Thus it seems that these marine mammal tissues from different parts of the ocean have a more or less constant pattern of relatively persistent CBs. This would suggest that the CB patterns of persistent CBs in the different food sources are also very similar. Their concentrations in mammalian tissues appear to depend on the levels in their food sources and factors like sex and age.

#### Pesticides

In order to detect significant differences in absolute concentrations of individual pesticides between tissues within each of the groups of animals, a one-way analysis of variance was carried out ( $\alpha=0.05$ ). The results are given in Table 11. The average

**Table 9.** Averages and relative standard deviations (% of average, std%) of mol per cent contributions of individual CBs to their sum in each of the tissues of animals 1-11 (group A) and 12-19 (group B). For numbering of animals see Table 1 and of CBs Table 2. No entry: data do not allow a reliable quantitation. Significance levels for differences between groups A and B (two-tailed *t*-test): \**p*<0.05; \*\**p*<0.01; \*\*\**p*<0.001. Group I and group II congeners are separated by an empty row

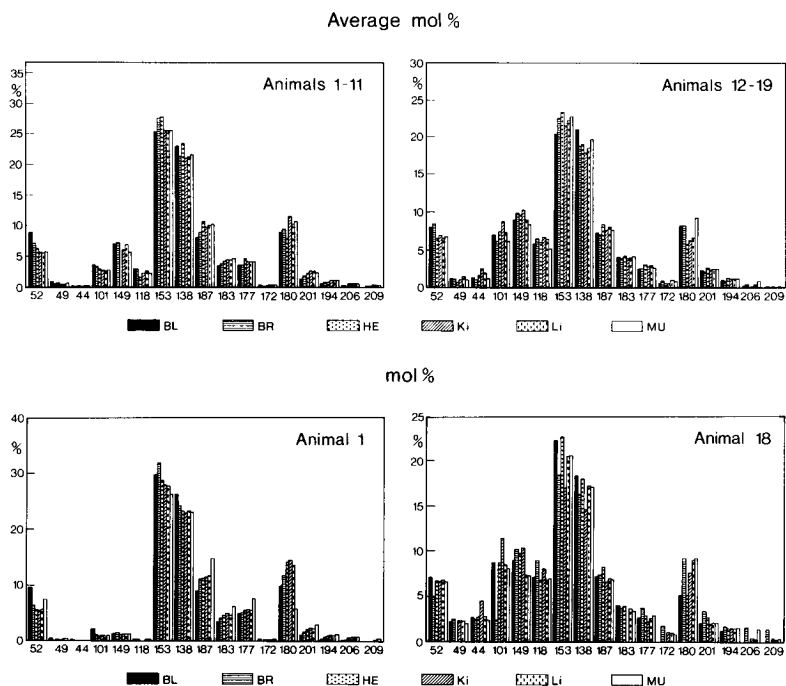
Animal group	Blubber				Brain				Heart			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
CBs												
52	9.0	25	8.1	13	7.2	30	8.4	18	6.3	17	6.5	8
49	1.0	70	1.2	89	0.5	122	1.0	130	0.8	64	0.6	141
44	0.4	96	1.3	89	0.2	166	1.0	131	0.3	81	1.6	72
101	3.7	58	6.9	28	3.5	53	6.0	33	3.0	62	7.3	19
149	7.1	57	8.9	11	7.4	47	9.8	15	4.3	80	9.6	6
118	2.9	62	5.9	32	3.0	53	6.4	41	1.7	97	6.0	17
153	25.4	9	20.3	27	27.5	12	22.4	14	27.8	3	23.3	3
138	23.0	7	21.0	15	21.3	10	18.7	13	23.5	4	18.9	4
187	8.1	10	7.4	14	9.0	12	7.1	9	10.7	5	8.3	5
183	3.6	12	4.1	13	3.9	17	3.9	13	4.4	7	4.1	8
177	3.6	23	2.5	30	3.7	19	2.6	31	4.7	14	3.1	7
172	0.5	40	0.6	61	0.3	110	0.9	57	0.3	39	0.6	72
180	9.1	31	8.2	40	9.4	43	8.2	33	8.0	63	5.9	8
201	1.4	28	2.3	42	1.9	36	2.1	35	2.4	27	2.7	14
194	0.7	47	1.1	12	0.9	41	0.9	63	1.0	12	1.4	7
206	0.3	111	0.3	134	0.2	190	0.5	119	0.7	71	0.0	
209	0.1	204	0.1	165	0.1	157	0.2	214	0.2	173	0.0	
Animal group	Kidney				Liver				Muscle			
	A		B		A		B		A		B	
	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%	avg	std%
CBs												
52	5.8	27	6.8	1	5.5	29	6.5	3	5.8	27	6.7	14
49	0.6	71	0.9	103	0.6	70	1.4	71	0.7	57	0.9	95
44	0.2	110	2.4	77	0.3	112	1.8	71	0.3	96	1.1	73
101	2.8	59	8.6	27	2.7	61	7.3	20	2.9	54	6.1	24
149	6.2	64	10.1	1	7.0	57	8.9	13	5.8	63	8.3	13
118	2.3	69	6.7	22	2.7	82	6.4	11	2.3	63	5.2	32
153	25.5	11	21.4	15	25.6	10	22.1	6	25.6	10	22.6	9
138	21.0	9	17.8	14	21.2	9	18.5	5	21.7	10	19.6	9
187	9.8	12	7.6	11	10.0	10	8.0	9	10.1	18	7.6	12
183	4.5	20	3.9	11	4.5	15	4.0	7	4.7	21	4.1	10
177	4.2	16	2.9	18	4.2	14	3.0	10	4.2	32	2.6	32
172	0.5	56	0.6	73	0.5	44	1.0	15	0.6	36	0.8	21
180	11.5	42	6.2	20	10.4	53	6.7	28	10.7	38	9.3	33
201	2.7	37	2.3	9	2.7	34	2.5	17	2.4	38	2.6	31
194	1.2	43	1.2	18	1.2	43	1.3	14	1.3	40	1.3	11
206	0.7	45	0.2	141	0.7	49	0.5	105	0.7	53	1.0	53
209	0.3	80	0.1	141	0.3	64	0.1	141	0.3	70	0.2	106

concentrations of most pesticides in group A are not significantly different between organs of animals of group A, with the exception of  $\alpha$ -HCH in brain, HCB

in brain and liver, DDD in brain and liver and DDT in blubber compared to all other tissues. In group B animals, differences in average concentrations occur

Table 9. Continued

CBs	Ratio of averages A/B					
	Blubber	Brain	Heart	Kidney	Liver	Muscle
52	1.1	0.9	1.0	0.8	0.8	0.9
49	0.9	0.5	1.5	0.7	0.5	0.8
44	0.3	0.2	0.2*	0.1*	0.1*	0.2*
101	0.5**	0.6**	0.4***	0.3***	0.4***	0.5***
149	0.8	0.8	0.5***	0.6**	0.8	0.7
118	0.5*	0.5**	0.3***	0.4***	0.4***	0.5**
153	1.2*	1.2**	1.2***	1.2**	1.2**	1.1*
138	1.1	1.1*	1.2***	1.2**	1.1*	1.1*
187	1.1	1.3***	1.3***	1.3***	1.3***	1.3**
183	0.9*	1.0	1.1	1.2	1.1*	1.2
177	1.4**	1.4**	1.5***	1.5***	1.4***	1.6**
172	0.8	0.3**	0.5	0.7	0.5	0.7*
180	1.1	1.1	1.4	1.9**	1.5	1.2
201	0.6*	0.9	0.9	1.2	1.1	0.9
194	0.7*	1.0	0.7**	0.9	0.9	1.0
206	1.0	0.4		4.6	1.4	0.7
209	1.6	0.6		2.2	3.1	2.0



**Figure 8.** Mol % contributions (based on concentrations in  $\text{nmol g}^{-1}$  hexane extractable lipid) of CB congeners (identified by numbers, Table 2) to their mixtures in blubber (BL), brain (BR), heart (HE), kidney (KI), liver (LI) and muscle (MU) in animals 1–11 (group A, left) and animals 12–19 (group B, right). Y-axes have different scales. Top: average values of groups A and B. Bottom: values for individual animals 1 and 18.

**Table 10.** Results of a two-factor analysis of variance (ANOVA) involving the potential systematic influences of the differences between the two groups of animals (A and B) and the differences between the organs, on the relative concentrations of the individual pesticides and CBs. Concentrations were expressed as ratios to DDE and CB No. 153 (X/DDE and X/153, resp.). The underlying significance level was set at  $\alpha = 0.05$ . The F tests show whether the null-hypothesis of equality of means can be rejected. This is the case when the p values of the marginal F distributions are smaller than  $\alpha = 0.05$ . In those cases, significant differences are present. Given are the p values for the main (overall) effect and the tests of the null-hypothesis of equality in the two groups and in the organs. The ratios are statistically not significantly different for  $p > \alpha$

Compounds	p values		
	Groups of animals	Organs	Overall effect
<b>Pesticides</b>			
$\gamma$ -HCH	0.00	0.97	0.00
$\alpha$ -HCH	0.01	0.00	0.00
HCB	0.00	0.00	0.00
Dieldrin	0.00	0.74	0.03
DDD	0.00	0.28	0.00
DDT	0.00	0.00	0.00
<b>Chlorobiphenyls</b>			
52	0.01	0.01	0.00
49	0.01	0.73	0.16
44	0.00	0.70	0.00
101	0.00	0.40	0.00
149	0.00	0.49	0.00
118	0.00	0.60	0.00
138	0.18	0.19	0.15
187	0.85	0.59	0.69
183	0.05	0.70	0.39
177	0.00	0.43	0.02
172	0.00	0.76	0.00
180	0.94	0.74	0.84
201	0.00	0.12	0.12
194	0.00	0.32	0.06
206	0.47	0.03	0.04
209	0.47	0.38	0.40

for  $\alpha$ -HCH and  $\gamma$ -HCH in brain, for dieldrin and DDD between brain and muscle, for DDT in blubber and for DDE compared to blubber and muscle.

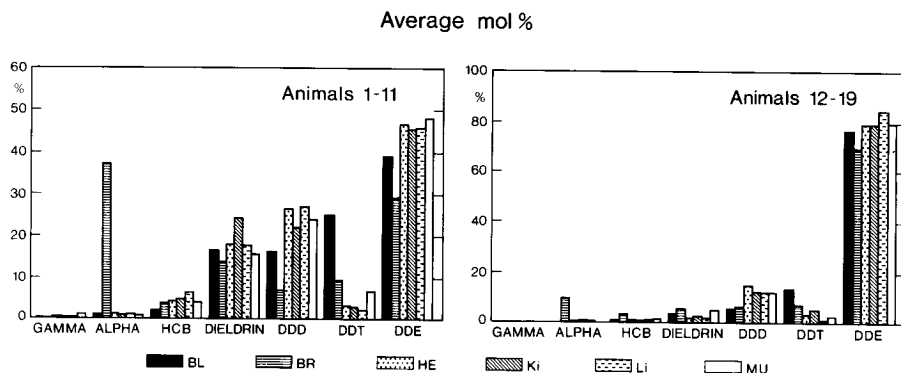
We shall use concentration data measured in blubber, brain, heart, kidney, liver and muscle of harbour porpoises and species of the open sea, in blood of seals fed with Atlantic fish (Atlantic group), Atlantic fish (mackerel) and in Wadden Sea fish. The latter were sampled at 17 stations in the Dutch and German Wadden Sea during two different periods in 1979. Whole fish (plaice and flounder) and

homogenates of shrimps were used for analysis. Contributions in shrimp and Wadden Sea fish are very similar. The composition of the pesticides mixture differed between the Wadden Sea and Atlantic fish (Fig. 12). The highest contributions in the Wadden Sea fish are those of HCB and DDD (40 and 30 mol % in the mixtures, reflecting the relative concentrations of these pesticides in Dutch coastal waters (Duinker & Hillebrand, 1979a). DDE and DDT were the dominant contributors in the Atlantic fish (45 and 20 mol %, respectively, Fig. 12).

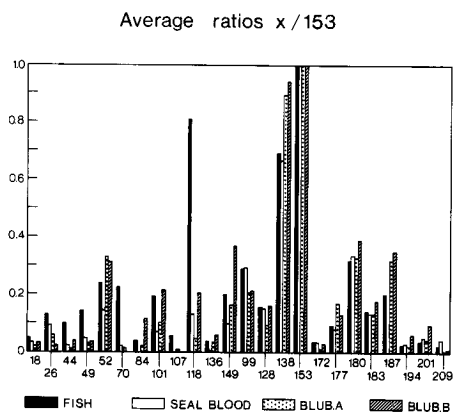
Comparison of the compositions of the pesticide mixtures (Fig. 12) and the individual concentrations in fish, blubber and liver in the coastal samples (Fig. 13) show that  $\gamma$ -HCH,  $\alpha$ -HCH and HCB are not biomagnified in liver and blubber. Figs. 12 and 13 summarize also the corresponding data for the Atlantic group of mammals.  $\gamma$ -HCH,  $\alpha$ -HCH, HCB, dieldrin, DDD and DDT occur at lower relative concentrations, while DDE is the dominant contributor to the mixtures (up to 85% in liver, Fig. 12). This high contribution is larger than in most other data reported (see Aguilar, 1985). Like in the coastal species,  $\gamma$ -HCH,  $\alpha$ -HCH and HCB are not biomagnified in blubber and liver. DDD and DDE are biomagnified in blubber and liver and DDT also in blubber but not in liver.

In the literature, the contributions of  $p,p'$ -DDT,  $p,p'$ -DDD and  $p,p'$ -DDE have often been added and described as  $\Sigma$ -DDTs. Figure 12 shows that the relative contributions of each compound to the mixture may vary considerably between sample types. The increase in the mol % contribution of DDE to  $\Sigma$ -DDT in open sea cetaceans in the series food < blubber < liver (up to 85%) parallels the porpoise data, but the relative contributions of DDE in the porpoise data are considerably smaller for all sample types (Fig. 12). The mol % contributions of DDD are larger in each of the corresponding coastal or Wadden Sea animals. A different situation occurs for DDT, where mol % contributions are much lower in Wadden Sea fish than in Atlantic fish, although relative contributions to blubber and liver of cetaceans are larger. Comparison of average concentrations in blubber and liver of the two groups of animals A and B show that the ratio:— $\{\text{blubber A/liver A}\}/\{\text{blubber B/liver B}\}$  is very similar for DDD, DDT and DDE, suggesting that similar mechanisms may be operative in the animal groups A and B.

The relatively high concentration of  $\alpha$ -HCH in brain tissue compared to other compartments in especially porpoises has also been reported for Wadden Sea seals (Duinker & Hillebrand, 1979b). For all other—less polar—compounds, brain levels are the lowest. An hematoencephalic barrier was assumed to be responsible for this phenomenon (Franck, Ronald & Braun, 1973), but lipid com-



**Figure 9.** Average mol % contributions (based on concentrations in  $\text{nmol g}^{-1}$  hexane extractable lipid) of  $\gamma$ -HCH (GAMMA),  $\alpha$ -HCH (ALPHA), HCB, dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) to their mixtures in blubber (BL), brain (BR), heart (HE), kidney (KI), liver (LI) and muscle (MU) in animals 1-11 (group A, left) and animals 12-19 (group B, right). Y-axes have different scales.



**Figure 10.** Average concentration ratios ( $X/153$ ) for CB congener X (identified by numbers, Table 2) relative to CB-153 in Wadden Sea fish, seal blood, blubber (BLUB) of North Sea porpoises of group A (BLUB.A) and blubber of the species of group B (BLUB.B). The order of CBs on the x-axis is different from that in the preceding graphs. They have been arranged in the order of increasing CB number for both group I congeners (CB-18 through CB-149) and group II congeners (CB-153 through CB-209).

position was considered to be a more plausible explanation by others (Roberts, de Freitas & Gidney, 1977; Mitchell, Plack & Thompsom, 1977). The large fraction of relatively polar phospholipids in brain would account for the larger concentrations of relatively polar compounds such as  $\alpha$ -HCH (Kawai *et al.*, 1988).

Since e.g., CBs show the same patterns in all tissues, including brain, the present results would support the latter hypothesis. The distribution patterns of pesticides in different tissues and organs

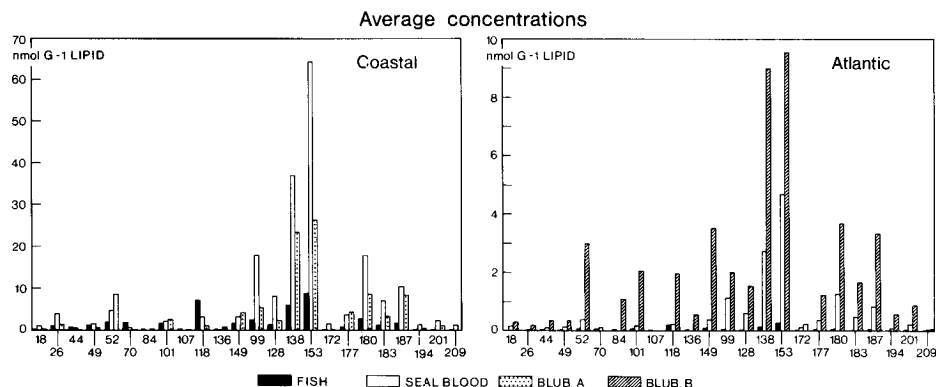
of animals from the two groups are considerably more variable than those of the CBs. This reflects the different chemical nature of the pesticides.

In order to test any difference in affinity of the pesticides for specific tissue lipids, a one-way analysis of variance was carried out, involving concentration ratios blubber/brain, blubber/heart, etc., within each of the groups A and B. The underlying significance level was set at  $\alpha=0.05$ . The results are given in Table 12. Significant differences occur in the blubber/brain ratios for  $\gamma$ -HCH, HCB, DDD and DDE relative to all other ratios and for  $\alpha$ -HCH relative to blubber/kidney in group A. These differences are reflected in the average values as shown in Figure 7. Average blubber/brain ratios in group B are different from all other ratios for  $\gamma$ -HCH, dieldrin and DDD, more complicated relations occur for  $\alpha$ -HCH and HCB. For DDE, differences between tissues are not significant, reflecting the large standard deviations (Table 8).

Without more fundamental studies on the kinetics of pesticides in mammals being available, no more quantitative relations can be established from the present data.

#### Lipid compositions

The concentrations of congeners on a wet weight basis vary considerably between tissues of one animal. The largest part of these differences disappear when concentrations are expressed on a hexane-extractable lipid basis (Table 4, Fig. 4), but smaller differences remain. Especially levels in brain appear to be considerably lower, except for  $\alpha$ -HCH, which showed considerably higher lipid based concentrations in brain compared to all other tissues. This was also reported by Kawai *et al.* (1988). These authors also investigated the types of lipid classes present in different cetacean tissues. In brain the total lipid fraction mainly consisted of phospholipids



**Figure 11.** Average concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of CBs (identified by numbers, Table 2) in left: Wadden Sea plaice (FISH), blood of seals fed with Wadden Sea fish (SEAL BLOOD) and blubber of North Sea porpoises (BLUB. A); right: Atlantic fish (mackerel, FISH), blood of seals fed with Atlantic fish (SEAL BLOOD) and blubber of other species (group B animals, BLUB. B). The y-axes have different scales.

(50–70%) and total cholesterol (15–30%). Triacylglycerols constituted only a minimal fraction in brain and increased in the series <liver, heart < muscle < kidney < blubber, where triacylglycerols were the only lipid class present. In general, the triacylglycerol fraction increased with increasing amounts of total extractable lipids of the different organs, but brain formed an exception to this rule in that it was relatively lipid rich, but the contribution of triacylglycerols was minimal. In conclusion, organochlorines appear to follow the basic chemical rule that a given compound dissolves best in a 'solvent' of similar polarity.

The method involving the use of an organic solvent (or a mixture of solvents) to estimate the amount of 'extractable' lipids is an admittedly inadequate way to represent the 'site' of CBs in animal tissues, taking into account the complex nature of lipids and the different specific affinities of individual compounds for different lipid constituents and other apolar moieties (Kawai and Fukushima 1981). For instance, Tanabe *et al.* (1988) argued that the content of the sum of triacylglycerols and free fatty acids in a tissue would be a better basis for expressing the concentration of lipophilic compounds in that tissue. The large contribution of waxes to blubber, muscle and organs of the sperm whale may partly account for the low extractable lipid-based concentrations of all CBs in all tissues of animal 15, but other factors like feeding habits may also contribute.

The suggestion of Tanabe *et al.* (1988) is favourable to the standardisation of concentration levels on triacylglycerols alone, since blubber did not show higher OC concentrations than less total-extractable lipid- and triacylglycerol-rich tissues such as liver, muscle, kidney and heart in the present study. This observation suggests that the covariance of lipids and

chlorinated biphenyls is a function of specific affinity of CBs for certain types of lipids and not a non-specific affinity for all lipids and other hydrophobic moieties (Phillips, 1986).

We feel therefore that considerably more work is necessary to develop methods for more precise and accurate correlations with the actual apolar terminals in tissues which secure the lipophilic organochlorines being studied. In particular, it seems highly desirable to adopt a standardized procedure for analyzing a broad spectrum of lipid classes as has been performed in some cases (Roberts, de Freitas & Gidney, 1977; Mitchell, Plack & Thomsson, 1977; Tanabe *et al.*, 1981a,b; Schneider, 1980, 1982; Kawai *et al.*, 1988).

It was mentioned above that the mol % contributions of several pesticides differ considerably between tissues. These observations would result if lipids of the various tissues have larger affinity differences for pesticides than for CBs. The lipid composition of a given tissue type is expected to vary only slightly between different animals (Tanabe *et al.*, 1988). Therefore, each block of bars in Figure 2 (grouped per compound in each animal) would be expected to be qualitatively similar for each compound in all animals. The overall height of the block in this figure would vary between animals according to varying burdens of the compounds. Also, we would expect quite large differences between blocks (representing different pesticides). This expectations are confirmed to a large extent by the data (Figs. 2 and 9).

#### *Comparison between animals found dead and those kept in captivity*

Several authors have expressed serious concern whether the levels of OCs determined in organisms found dead can be considered as representative for



**Table 11.** Results of a ONE-WAY analysis of variance of the absolute (lipid based) concentrations of the pesticides and chlorobiphenyls in each of the two groups of animals (A and B), represented by the p values ( $\alpha=0.05$ ) in the columns 'ONE WAY'. The ratios are not significantly different for  $p > \alpha$ . Also a Tukey test (a modified *t*-test) was carried out. The significant differences of the average concentration of any compound between tissues are identified in the columns 'Tukey test'. If one tissue is given, its average concentration was significantly different from those of all other tissues (bl = blubber; br = brain; he = heart; ki = kidney; li = liver; mu = muscle). Expressions like 'br vs bl,li,mu' indicate that the average brain concentration was significantly different from those of blubber, liver and muscle (implying that brain, heart and kidney formed an homogeneous group, and similarly blubber, heart, kidney, liver and muscle formed another homogeneous group)

Compounds	Group A animals		Group B animals	
	ONE WAY p	Tukey test	ONE WAY p	Tukey test
<b>Pesticides</b>				
$\gamma$ -HCH	0.08	—	0.00	br
$\alpha$ -HCH	0.00	br	0.00	br
HCB	0.06	br vs li	0.03	—
Dieldrin	0.27	—	0.01	br vs mu
DDD	0.08	br vs li	0.00	br vs mu
DDT	0.00	bl	0.00	bl
DDE	0.39	—	0.00	br vs bl,mu
<b>CBs</b>				
52	0.04	br vs mu	0.03	—
49	0.00	br vs bl,li,mu	0.02	bl vs br
44	0.04	br vs mu	0.01	br vs bl,mu
101	0.00	br vs bl,li,mu	0.02	—
149	0.31	—	0.03	—
118	0.29	—	0.02	bl vs br
153	0.06	—	0.06	—
138	0.06	br vs mu	0.03	—
187	0.08	—	0.03	br vs mu
183	0.15	—	0.03	—
177	0.06	—	0.07	—
172	0.11	—	0.03	—
180	0.37	—	0.07	—
201	0.17	—	0.15	—
194	0.27	—	0.02	br vs mu
206	0.03	—	0.03	—
209	0.14	—	0.18	—

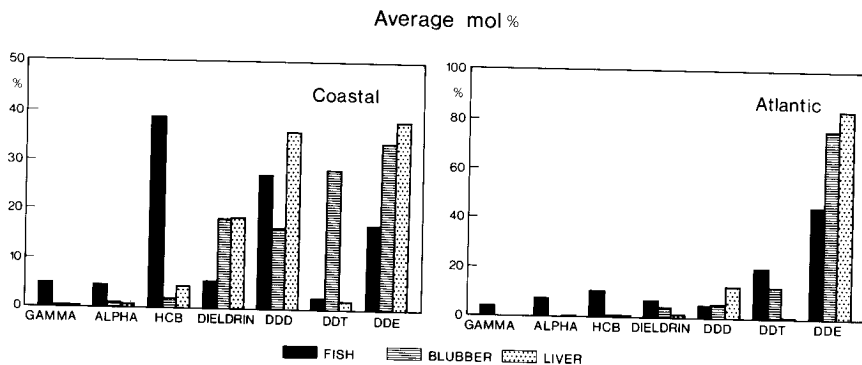
levels in the population (Aguilar, 1985; Olsson, 1986). Bergman, Olsson & Reuthergaardh (1981) found that average PCB concentrations in animals found dead were two times higher than those in actively killed ones. Stranded organisms may have suffered from diseases before death occurred and decay afterwards. The constancy of the composition of mixtures in different tissues of animals, sampled once found dead and of those which died at the Dolfinarium at Harderwijk, as well as the good correlation with seal blood taken from animals in captivity suggest that the problem, if existing, affects all CBs similarly. An important point might be that blood as transport compartment, stops flowing after death.

There are nevertheless several reasons (e.g., tissue inhomogeneities, differences in sample treatment, instrumental analytical and data reporting techniques) (Aguilar, 1985; Reuthergaard & Knap, 1987) why data on organochlorine compounds in marine mammal tissues reported by different authors may not be well comparable (Andersen & Rebsdorff, 1976; Gaskin, Holdrinet & Franck, 1976; Drescher, Harms & Huschenbeth, 1977; Donkin, Mann & Hamilton, 1981; Wagemann & Muir, 1981).

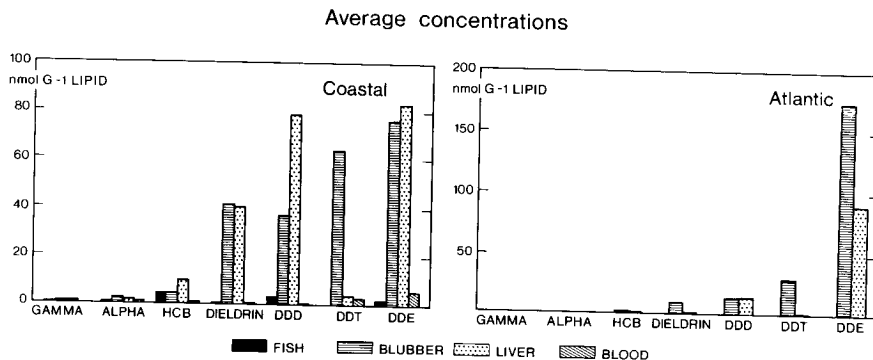
### Conclusions

#### *Chlorinated biphenyls*

Variable concentrations of individual chlorinated biphenyls (CBs) and pesticides were established in



**Figure 12.** Average mol % contributions of  $\gamma$ -HCH (GAMMA),  $\alpha$ -HCH (ALPHA), HCB, dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) to their mixtures in Wadden Sea fish (left) and Atlantic fish (right) and blubber and liver of animals 1-11 (North Sea porpoises, group A, left) and animals 12-19 (other species, group B, right).



**Figure 13.** Average concentrations ( $\text{nmol g}^{-1}$  hexane extractable lipid) of  $\gamma$ -HCH (GAMMA),  $\alpha$ -HCH (ALPHA), HCB, dieldrin, p,p'-DDD (DDD), p,p'-DDT (DDT) and p,p'-DDE (DDE) in fish, blubber, liver and blood. Left: Wadden Sea fish, and blubber, liver and blood of North Sea porpoises (group A, animals 1-11); Right: Atlantic fish, and blubber and liver of the other species (group B, animals 12-19).

different tissues of the same organism and between organisms from the coastal area, the open North Sea and the western Atlantic. Differences in lipid based concentration in the same animal were generally lower than between animals.

Maximum concentrations of CBs and pesticides were of the same order of magnitude ( $50\text{--}100 \text{ nmol g}^{-1}$  lipid with DDE values up to 500).  $\sum\text{-PCB}/\sum\text{-DDT}$  ratios which have been used frequently in the literature tend to obscure this observation: a value around 10 has been considered characteristic for several areas. Moreover, sums of concentrations do not reveal and distinguish the compositions of the mixtures. Significant differences in the relative amounts of p,p'-DDT, p,p'-DDD and p,p'-DDE have been established between tissues (large contributions of DDT in blubber and of DDE in liver and considerable differences between the porpoises and the open sea species) which make comparisons of their sums scientifically irrelevant.

The mixtures of persistent CBs appeared to be very similar in different mammal tissues from the Dutch coastal area, the North Sea, the western Atlantic and literature data from the Southern Indian Ocean. The mixtures of CBs in marine mammal tissues and organs are dominated by highly chlorinated biphenyls. This does not reflect the composition in ingested food, as was also suggested by Phillips (1986).

The differences in biomagnification can be interpreted in terms of the number and relative positions of chlorine atoms in the molecular framework, suggesting biotransformation to be the most probable cause for the decreased biomagnification of several congeners. The apparent contribution of biotransformation to the kinetics of PCBs in marine mammals was seals > porpoises > open sea cetaceans. The differences between these classes can be caused by physiological differences or a higher degree of mixed function oxygenase induction at

**Table 12.** Results of a ONE-WAY analysis of variance (ONE-WAY) of pesticide concentration ratios in the different organs of each of the two groups of animals A and B. bl/br = blubber/brain; bl/he = blubber/heart; bl/ki = blubber/kidney; bl/li = blubber/liver; bl/mu = blubber/muscle. The underlying significance level was set at  $\alpha=0.05$ . p values are given in the appropriate columns 'ONE WAY'. Also a Tukey test (a modified *t*-test) was carried out. In the corresponding columns, those ratios are identified that differed significantly from others; e.g., an entry bl/br indicates that the average blubber/brain ratio differed significantly from all other average ratios for that compound. Similarly, an entry 'bl/br vs bl/ki' indicates that the average blubber/brain ratio differed significantly from bl/ki, while average bl/br, bl/he, bl/li, bl/mu values and also the group of average bl/he, bl/ki, bl/li, and bl/mu values formed two homogeneous groups

Compounds	Group A animals		Group B animals	
	ONE WAY p	Tukey test	ONE WAY p	Tukey test
$\gamma$ -HCH	0.00	bl/br	0.00	bl/br
$\alpha$ -HCH	0.01	bl/br vs bl/ki	0.00	bl/br; and bl/he vs bl/ki and bl/ki vs bl/mu
HCB	0.00	bl/br	0.00	bl/br vs all but bl/ki
Dieldrin	0.05	—	0.00	bl/br
DDD	0.00	bl/br	0.00	bl/br
DDT	0.00	—	0.08	—
DDE	0.00	bl/br	0.36	—

higher levels of PCBs. CB-congeners with vicinal H-atoms only in the m,p position were apparently metabolized in seals, but were persistent in cetaceans. CB-congeners with vicinal H-atoms in the o,m position with one ortho-chlorine at maximum, were metabolized in each case. Although CBs with either vicinal H-atoms only in the m,p position or only in the o,m position with two ortho-chlorines were persistent in cetaceans, congeners possessing both features were apparently metabolized.

Concentrations in brain are lower than in all other organs investigated. Enriched brain concentrations of early eluting congeners (involving metabolizable ones), as found by Tanabe *et al.* (1988) were not found here.

The biomagnification and distribution of CBs is probably due to specific affinities of CBs for lipids rather than to a non-specific affinity of CBs for all lipids (Phillips, 1986). This is supported by the general finding that the lipid based concentrations of CBs decreased in the order brain < heart, kidney, liver < blubber and the decrease in biomagnification with increasing degree of chlorination in blubber of harbour porpoises.

No indication was found for accumulation in blubber at the expense of other tissues since patterns were virtually identical. Comparison with data on experiments with PCBs and seals (Boon *et al.*, 1987) shows that biomagnification of metabolizable congeners was lower in blood, suggesting that the uptake from food into blubber is more effective and/

or elimination from blubber is faster than for blood. Net transfer from food to faeces, blood and blubber appeared to be very similar for persistent congeners.

It is highly desirable that analyses of OCs be accompanied by lipid speciation studies, for which methods have been proposed in the literature. This would further increase the usefulness of CBs as model compounds for the behaviour of lipophilic organic compounds in the environment. It is therefore essential that in studies of PCBs and other mixtures of organochlorine compounds, individual compounds are studied, rather than undefined mixtures.

#### Pesticides

The patterns of pesticides are considerably more variable than those of CBs, reflecting the larger differences in molecular structures and properties. Relatively large differences were found in the compositions of the pesticide mixtures between tissues of the same animal and between those of the porpoises and the other species from the open North Sea and western Atlantic. DDE is the dominant contributor in the tissues of the latter group (up to 85% of the mixture in liver). Also in Atlantic fish (mackerel) DDE was found to be the dominant contributor, in contrast to the situation in Wadden Sea fish where DDD is the dominant contributor, reflecting the relative concentrations of DDT, DDT and DDE in solution in Dutch coastal waters (Duinker & Hillebrand, 1979a).

The concentrations of some OC pesticides in marine organisms appear to be similar to, or even larger than those of individual chlorobiphenyls. The popular method to express the relative concentrations of PCBs and DDTs as the ratio  $\sum \text{PCB} / \sum \text{DDT}$  (often found to be around 10) therefore underestimates the possible role of OC pesticides for marine mammals compared to CBs. We recommend that OC pesticides are analysed in marine mammal tissues, in addition to individual CBs.

$\alpha$ -HCH is a relatively polar compound. Its relative concentration is low in all organs, except in brain. This is more likely to result from the high fraction of the polar phospholipids in brain lipids than due to the presence of a hematoencephalic barrier.

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